

Review

# Functionalized Carbon Nanotubes (CNTs) for Water and Wastewater Treatment: Preparation to Application

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**Abstract:** As the world human population and industrialization keep growing, the water availability issue has forced scientists, engineers, and legislators of water supply industries to better manage water resources. Pollutant removals from wastewaters are crucial to ensure qualities of available water resources (including natural water bodies or reclaimed waters). Diverse techniques have been developed to deal with water quality concerns. Carbon based nanomaterials, especially carbon nanotubes (CNTs) with their high specific surface area and associated adsorption sites, have drawn a special focus in environmental applications, especially water and wastewater treatment. This critical review summarizes recent developments and adsorption behaviors of CNTs used to remove organics or heavy metal ions from contaminated waters via adsorption and inactivation of biological species associated with CNTs. Foci include CNTs synthesis, purification, and surface modifications or functionalization, followed by their characterization methods and the effect of water chemistry on adsorption capacities and removal mechanisms. Functionalized CNTs have been proven to be promising nanomaterials for the decontamination of waters due to their high adsorption capacity. However, most of the functional CNT applications are limited to lab-scale experiments only. Feasibility of their large-scale/industrial applications with cost-effective ways of synthesis and assessments of their toxicity with better simulating adsorption mechanisms still need to be studied.

**Keywords:** carbon nanotubes; surface modification; heavy metals; adsorption; water and wastewater treatment

## 1. Introduction

Rapid urbanization and industrialization has significantly increased the clean water demands in the domestic, industrial, and agricultural sectors [1–3]. Meanwhile, large quantities of pollutants including organic, inorganic, and biological contaminants are being released into the water bodies from these sectors [4–6]. Eccentric waters such as brackish, storm, and wastewater are being used depending upon the purposes [7,8]. Increasingly, use of these waters has also increased the urgent concern about the burden of negative impacts on the surrounding environment; one of the tremendous challenges confronting mankind is the exploration of green and sustainable methods to overcome these shortcomings [9–13]. Keeping in mind the current situation of water and wastewater treatment status, the technologies are not sustainable to meet healthy requirements for surrounding environment and community health [14,15].

Historically, numerous techniques and methods have been investigated for advanced treatment of water and wastewater [16]. The most common, adsorption, was proven to be the improved technique to remove a variety of pollutants including organic and inorganic contaminants present in water and wastewater [17,18]. Limited treatment efficiency was reported by using conventional adsorbents due to their small surface area, limited number of active sites, deficiency in selectivity, and low adsorption kinetics [19]. These shortcomings of conventional adsorbents have been addressed in recent advancements of nano-adsorbents owing to their high surface area coupled with a higher number of active sites, tunable pore size, fast kinetics, and improved surface chemistry [10,20–24].

The nanomaterials can be used for treatment of water and desalination as well and also reveal properties including electron affinity, mechanical strength, and flexibility during functionalization [25–27]. Carbon nanomaterials (CNs) such as carbon nanotubes (CNTs) are supposed to be a promising material to break down the tradeoff concerning selectivity and adsorption, resulting in an increase of the economics of adsorption technology [25]. As a result, CNTs, as an adsorbent for treatment of water, have attained the focus of countless scholars over the previous few decades who are projected to carry on the exploration and developments in the field of CNs [28].

Numerous significant articles have been published on nanomaterials applied for the treatment of water and wastewater in previous few years [29–37]. Despite rapid developments, innovations, and applications of CNT-based nanomaterials, there is an increasing need for an across-the-board review of the synthesis of CNTs, functionalization of surface modifications, and finally their application to remove aqueous contaminants and to identify potential directions. This is the main motivation of the current review article.

This review attempts to address a brief history of CNTs, synthesis, purification, and functionalization, followed by the application of these nanomaterials for eliminating organics, inorganics, and microorganisms present in water and wastewater samples.

### *1.1. Historical Background*

The discovery of CNTs was reviewed in 2006 by Monthieux and Kuznetsov, showing that the science has seemed to remain controversial [38]. Most literature mentions that nanotubes were discovered by Sumio Iijima [38]. However, Radushkevich and Lukyanovich explained the synthesis process of CNTs with 50 nm diameter [39]. Oberlin et al. explained the vapor phase growth technique for the synthesis of carbon fibers; the synthesized tubes consisted of turbostratic stacks of carbon layers (i.e., describing a crystal structure in which basal planes have slipped out of alignment) [40]. In addition, Abrahamson et al. [41] described the arc discharge method for carbon fiber synthesis using carbon anodes. Later on, scientists described the thermal catalytic disproportionation of CO for the synthesis of CNs. Transmission electron microscopy (TEM) and X-ray diffraction pattern (XRD) were used to characterize the synthesized CNs, and as a result they believed that CNTs can be formed by a graphene layer turning into a tubular shape. They also concluded that two types of promising arrangements, such as a helix-shaped spiral and circular arrangements in the form of a graphene hexagonal network, can result by turning the graphene layer into a tubular shape [42]. Later, a US patent was issued in 1987 on carbon nanofibers synthesis, the diameter ranging from 3.5 to 70 nm and five times greater in length than the diameter [43].

Back to the dates in 1950s, after the disclosure of CNTs by Iijima, projection of surprising properties of single-walled carbon nanotubes (SWCNTs) made by Dunlap and colleagues also attracted the attention of researchers around the world. At this time, after the discoveries and exploration of SWCNTs by Bethune and Iijima independently at IBM (Shiba, Minato) and Nippon Electric Co., Ltd. (Tokyo, Japan), respectively, the research on CNs and their specific methods of production was extended [44,45].

The above findings seem to be the extension of Fullerenes' discovery. Arc discharge technology had previously been applied for the production of laboratory-scale Buckminster

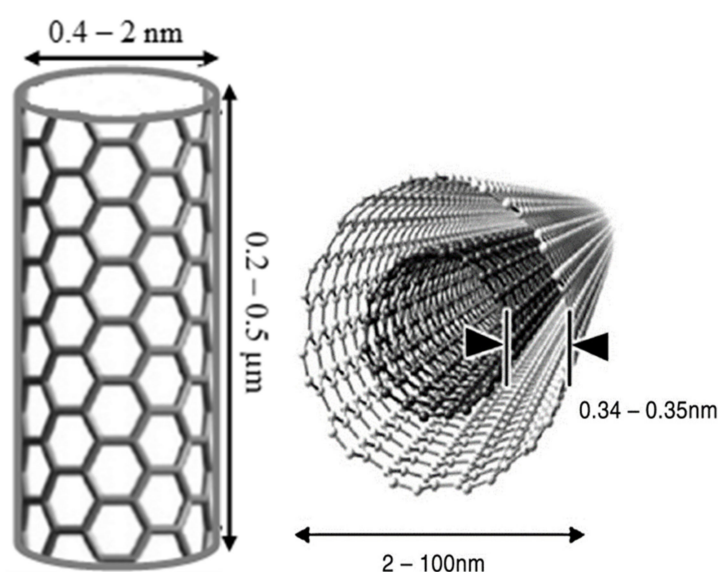
fullerenes [46,47]. CNTs are being studied after the report published in 1991 by Iijima [45] was fundamental, because it put CNTs in the limelight [38].

### 1.2. Types of CNTs and Structure

CNTs are composed of carbon atoms organized in a progression of fused benzene rings, which are pleated into a cylindrical shape. This new sort of man-made nano-material has a place with fullerene family and is treated as carbon's third allotrope as well as  $sp^2$  and  $sp^3$  forms of graphite and diamond, respectively [42,48,49].

Generally, there are two types of CNTs [50] on the bases of number of layers shown in Figure 1:

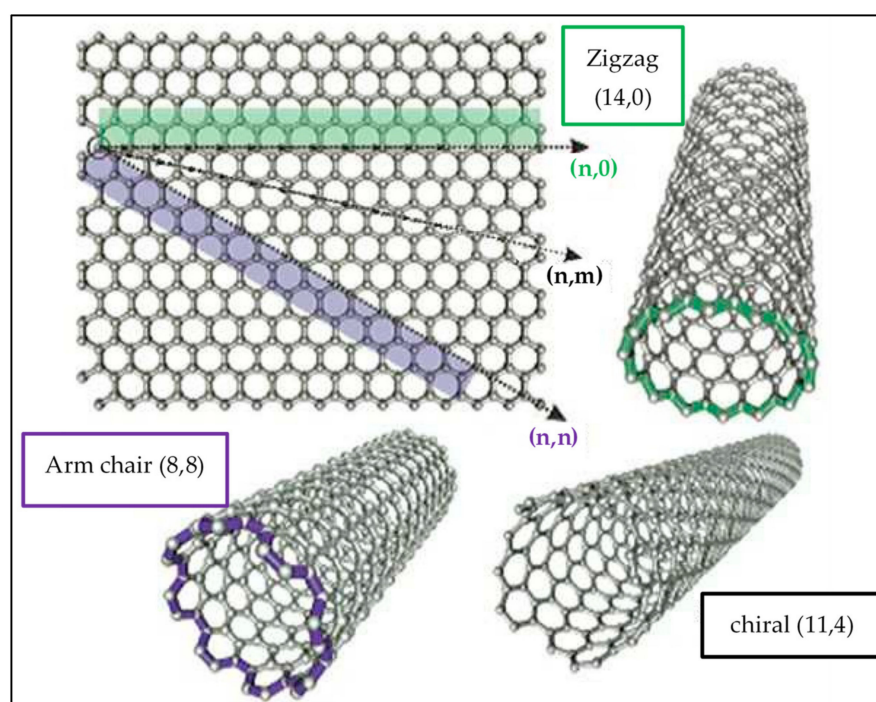
1. Single-walled carbon nanotubes (SWCNTs)
2. Multi-walled carbon nanotubes (MWCNTs).



**Figure 1.** Types of typical dimensions of CNTs, SWCNTs (left) and MWCNTs (right).

CNTs are made by a sheet of graphene when rolled into a cylindrical shape, which may have a capped or open end, usually in a hexagonal form close packed with a diameter at a small scale of 1 nm, and a few microns long. SWCNTs (Figure 1) with a diameter as small as 0.4 to 2 nm are made by the single sheet of graphene rolled into a cylindrical shape, while MWCNTs (Figure 1) with an outer and inner diameter ranging from 2–100 nm and 1–3 nm, respectively, and a several microns in length are made up of two or more sheets of graphene incasing a hollow core in the same way as in SWCNTs [49,51,52].

Based on the chemistry, there are two zones of CNTs: the sidewall and tip. A significant aspect in controlling these distinctive properties emanates after the change in the tube-like structure due to entrapment of graphene layers into a cylindrical shape. Figure 2 shows different structures of rolled SWCNTs based on graphene sheets. Depending on alignment of the cylinder axis relative to the hexagonal matrix, the CNTs structure can be stipulated by chiral carrier in three ways, armchair, chiral, and zigzag, illustrated by their chirality index ( $n,m$ ). Geometric arrangement of carbon atoms present at the layer of nanotubes is responsible for the foundation of zigzag ( $m = 0$ ) and armchair ( $n = m$ ) CNTs, whereas the structure of the nanotube with the two enantiomorphs on the right side is chiral ( $n \neq m$ ) [53,54].



**Figure 2.** Roll-up of graphene sheet into different types of CNTs [54,55].

Recent reviews are good to find detailed elucidations of the structure of CNTs [48,49,56–61]. Here, Table 1 summarizes a comparison of the properties of SWCNTs and MWCNTs.

**Table 1.** Comparison between properties of SWCNTs and MWCNTs [49,62,63].

Properties	SWNTs	MWCNTs
Layer type	Single graphene layer	Multiple graphene layer
Catalyst requirement	Essential during synthesis	No need during synthesis
Bulk or massive production	Difficult	Easy
Purity level	Low	Large
Defect's level	High	Low
Characterization	Easy	Difficult
Manage	Easily twisted	Cannot be twisted easily
Specific gravity	About 0.8 g/cm <sup>3</sup>	Less than 1.8 g/cm <sup>3</sup>
Elastic modulus	About 1.4 TPa	Ranging from 0.3 to 1 TPa
Strength	Ranging from 50 to 500 GPa	Ranging from 10 to 60 GPa
Electrical conductivity	Ranging from 102 to 106 S/cm	Ranging from 103 to 105 S/cm
Electron mobility	About 105 cm <sup>2</sup> /(V s)	Ranging from 104 to 105 cm <sup>2</sup> /(V s)
Thermal conductivity	About 6000 W/(m K)	About 2000 W/(m K)
Coefficient of thermal expansion	Greater than $1.1 \times 10^{-3} \text{ K}^{-1}$	About $-1.37 \times 10^{-3} \text{ K}^{-1}$
Thermal stability in air	Ranging from 600 to 800 °C	Ranging from 600 to 800 °C
Resistivity	Ranging from 50 to 500 $\mu\Omega \text{ cm}$	Ranging from 50 to 500 $\mu\Omega \text{ cm}$
Specific Surface Area	Ranging from 400 to 900 m <sup>2</sup> /g	Ranging from 200 to 400 m <sup>2</sup> /g

## 2. Synthesis of CNTs

Typically, there are three extensive methods for the synthesis of CNTs as given bellow:

- Arc discharge
- Laser ablation
- Chemical vapor deposition (CVD).

CNTs are produced by using energy and carbon sources in all the synthesis methods. A carbon electrode or a gas and an electric current or heat is used as a carbon and energy source, while using arc discharge or CVD methods, respectively, for the synthesis of CNTs,



whereas a laser beam is used as an energy source during the laser ablation method. Table 2 presents a detailed summary of the efficiencies of the CNT synthesis methods. These methods are based on the formation of a single or a consortium of carbon atoms that sack to recombine into CNTs.

CNT synthesis mechanisms have been debated in detail by Cassell et al. [64] and Sinnott et al. [65]. It is believed that by using the metallic catalyst in the CVD method for CNT synthesis, the cylindrical shaped graphene tube is formed by initial deposition of carbon atoms on the used catalyst surface [66]. It was also concluded that the size of particles of the used catalyst also play an important role in the CNTs diameter, as the catalyst particles in a smaller size produce SWCNTs with a diameter of a few nanometers, while the larger particles tend to produce MWCNTs [67].

The arc discharge method between graphite electrodes is the first method of producing CNTs [68]. Briefly, in this method, direct current of 50 to 100 A and about 20 volts of potential difference is established between a graphite electrode pair in the presence of one of the inert gases containing helium or argon with a pressure of 500 Torr [69,70]. The carbon electrode surface evaporates and forms a cylindrical-shaped tube structure because of the high temperature generated due to the discharge of electric current in low pressure, inert gas, and catalyst [44,70,71]. MWCNTs can be synthesized via an arc discharge method without a metallic catalyst; on the other hand, mixed-metal catalysts, for example iron, cobalt, and nickel, are necessary for SWCNTs fabrication [72]. In general, higher levels of structural precision are noted in CNTs produced via the arc discharge method [73]; however, different variables such as chamber temperature, concentration and type of catalyst, hydrogen presence, etc., may affect the structure and size of synthesized CNTs [74]. Recently, nickel-filled CNTs were synthesized via a local arc discharge method in liquid ethanol [75], nitrogen-doped CNTs via vaporization of boron nitride [76], low-cost SWCNTs via an arc discharge method in open air [77], and SWCNTs and MWCNTs via a hot plasma arc discharge method [78].

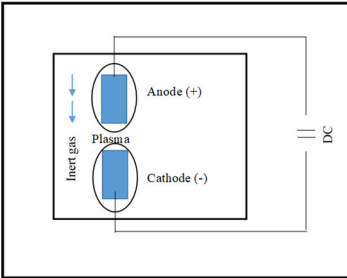
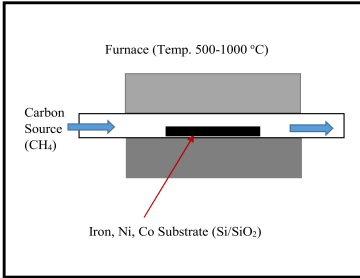
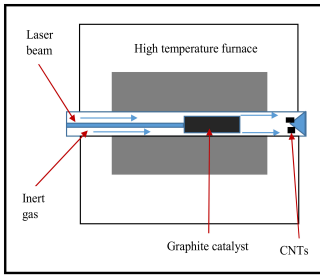
The use of laser ablation to synthesize CNTs was first reported by Guo and colleagues in 1995 [79,80]. Briefly, a graphite object is targeted by a laser beam with high energy in the presence of argon at 800–1200 °C temperature and 500 Torr pressure [69,70,81,82]. In this method, a laser pulse provides an energy source, and a graphite object serves as a source of carbon. Soot deposition of carbon can be avoided by uniform evaporation of the target resulting because of continuous applications of laser pulses. The larger size particles after the first laser beam are broken down into smaller ones by the successive beams. Later, the smaller size particles are produced into the CNT structure. Commonly, transition metals are used as catalysts in this method. Rope-shaped CNTs can be found by using a laser ablation method with the diameter ranging from 10 to 20 nm and about 100 mm long [83].

CVD is a well-liked method for bulk fabrication of CNTs around the globe. Typically, in this method, carbon monoxide or hydrocarbons gases are used as a source of carbon, while 500 to 1100 °C of temperature is used as the source of energy. The carbon atoms are deposited by the decomposition of the used carbon source and shaped into CNTs [84,85]. Briefly, the decomposition of gas (carbon source) occurs because of high temperature when transferred into reaction chamber together with the carrier gas and generates a substrate of carbon atoms on the surface of catalyst to form CNTs [86]. When compared to other methods of synthesis of CNTs, the CVD is the most common route for relatively large-scale production of CNTs as it is simple in operation, higher in yield, and economic and has a high rate of deposition and good control over the morphology of tubes during the synthesis process [87–89]. Cassell et al. [64] studied that CNTs in bulk can be produced, especially SWCNTs via the CVD method, by using acetylene as a source of carbon deposition in the presence of iron and cobalt and zeolite or silica as a carrier support material. They also concluded that SWCNTs can be produced on a largescale when a mixture of H<sub>2</sub> and CH<sub>4</sub> is deposited on the catalyst (Co or Ni), and MgO as carrier support material is used.

**Table 2.** A summary of CNTs synthesis strategies and their efficiencies.

Parameters		Arc Discharge Method	Chemical Vapor Deposition	Laser Ablation (Vaporization)	Ref.
Method	Source of energy	Direct current	Temperature (ignition)	High intensity laser beam	<a href="#">[90,91]</a>
	Source of carbon	Carbon or graphite electrodes	Hydrocarbon gases or carbon monoxide (CH <sub>4</sub> , CO, or acetylene)	Graphite object	
Temperature (°C)		3000 to 4000	500 to 1100	About 3000	<a href="#">[84,85]</a>
Cost per unit synthesis		Costly	Economic	Costly	<a href="#">[83,92]</a>
CNTs selectivity		Less	High	Less	<a href="#">[93]</a>
Availability of carbon source		Complex	Easy	Difficult	<a href="#">[94]</a>
Purification level		More	Less	More	<a href="#">[95]</a>
Nature of synthesis process		Batch	Continuous	Batch	<a href="#">[64,96]</a>
Control on synthesis parameters		Difficult	Easy	Difficult	<a href="#">[97,98]</a>
Energy requirement		High	Low	High	<a href="#">[99]</a>
Design of reactor		Hard	Simple and easy	Hard	<a href="#">[100]</a>
Nanotube graphitization		High	Moderate	High	<a href="#">[101–103]</a>
Typical yield		30 to 90%	20 to 100%	Up to 70%	<a href="#">[84,91,95,104–106]</a>
Typical Diameter	SWCNTs	0.6 to 1.4 nm	0.6 to 4 nm	1 to 2 nm	<a href="#">[87,91,107,108]</a>
	MWCNTs	Inner: 1 to 3 nm Outer: ~10 nm	0.1 to several nanometers	10 to 20 nm	<a href="#">[79,83,91,104,109]</a>

Table 2. Cont.

Parameters	Arc Discharge Method	Chemical Vapor Deposition	Laser Ablation (Vaporization)	Ref.
Advantages	<ol style="list-style-type: none"> <li>1. Synthesis of both SWCNTs and MWCNTs is easy</li> <li>2. MWCNTs can be produced without any catalyst</li> <li>3. Costly process but less than laser ablation method</li> <li>4. Synthesis of CNTs is possible in open air</li> <li>5. High degree of structural perfection</li> </ol>	<ol style="list-style-type: none"> <li>1. Bulk production is easy</li> <li>2. More extensive length CNTs than other methods</li> <li>3. Simple and easy process</li> <li>4. Quite pure</li> <li>5. Alignment of produced CNTs is good</li> <li>6. Diameter and number of layers can be controlled</li> </ol>	<ol style="list-style-type: none"> <li>1. Primarily for SWCNTs</li> <li>2. Diameter of CNTs can be controlled</li> <li>3. Lower numbers of defects</li> <li>4. High degree of structural perfection</li> <li>5. Tubes' length can vary from 5 to 20 mm</li> </ol>	[98,110–113]
Disadvantages	<ol style="list-style-type: none"> <li>1. Received with some structural defects</li> <li>2. Short and randomly distributed in length and direction</li> <li>3. Lot of structural purification is needed</li> <li>4. Contains carbon impurities</li> </ol>	<ol style="list-style-type: none"> <li>1. Only used to produce MWCNTs</li> <li>2. Higher structural defect density</li> </ol>	<ol style="list-style-type: none"> <li>1. Costly technique due to expensive lasers beams</li> <li>3. Power needs are high</li> <li>4. Low yield</li> </ol>	[65,86,98,109]
Figures				[114]

### 3. CNT Purifications

Some of the impurities include an amorphous phase of carbon, particles of particular metals, or any other carrier material associated with CNTs that will have an effect on their execution performance [115]. Some typical purification technologies and their characteristics are discussed in Table 3. On average, the CNTs synthesized via the CVD method showed a purity level ranging from 5% to 10% [116]. Therefore, a broad purification of CNTs is necessary before being used for different applications. The detection and identification of different impurities associated with CNTs using different techniques have been discussed in Table 4. It is believed that the CNT structure may be affected to some extent when removing impurities, so there is always need for a compromise with the final structure after purification process [116]. The common CNTs purification methods are discussed below:

1. Oxidation
2. Acid treatment
3. Surfactant based sonication.

Oxidation is a decent manner to remove carbon [117–122] and metal [117,121,123–126] impurities associated with CNTs. One of the main shortcomings that occur using this process of purification is the oxidation of CNTs themselves along with the impurities, but fortunately, the loss of CNTs is smaller than the impurities [114]. The reason to oxidize these impurities is more defects or open structures associated with them. This is another reason that the attachment of these impurities is often observed with a used metallic catalyst, and this metal catalyst may also play a role in oxidation [117,118,123,124,127]. There are some factors, such as type of oxidant, time of oxidation, temperature, metal contents, and environment, which can affect the oxidation efficiency and final yield.

Typically, the method of acid treatment is used to eliminate the metallic impurities associated with CNTs. First of all, by oxidizing or sonication of the CNTs, the surface of associated impurities (metals) is made apparent to acid until the solvation and finally CNTs collect in suspension. It has been observed in a number of studies that by using HNO<sub>3</sub> for the purpose of acid treatment, it only affects the metallic impurities rather than the CNTs or other carbon containing impurities [117,118,122,125]. By using HCl for this purpose, the impacts on CNTs and carbon impurities are also observed to a small extent [117,123,126]. Acid treatment for purification of CNTs in diluted form (4M HCl) can show same results as by the HNO<sub>3</sub>, but the metal surface must be apparent to the applied acid to make the solvation [128].

Although purified CNTs are produced relatively by acid reflux, the nanotubes amalgamate, and the impurities that they capture are very hard or sometime impossible to remove by filtration [129]. Therefore, a surfactant-based sonication process is implemented generally by dissolving sodium dodecyl benzene sulfate (SDBS) in ethyl or methyl alcohol solution for this purpose. Since after sonication CNTs took longer to settle down, ultrafiltration is required and then annealed at a high temperature (about 1000 °C) in the presence of N<sub>2</sub> for 4 h. Annealing of CNTs is performed to optimize their structure. Surfactant-based sonication has been presented to be an effective method for removing tangled impurities associated with amalgamated CNTs [116].



**Table 3.** Typical purification technologies and their characteristics [130,131].

Technologies	Methods	Characteristics	
		Advantages	Limitations
Physical method	Filtration	1. Non-destructive	1. Not very effective 2. CNT samples need to be extremely dispersed 3. Purification of samples can be done in a limited quantity at a time
	Centrifugation	2. Retains the inherency and intrinsic structure necessary to elucidate the properties of CNTs	
	Solubilization with functional groups	3. More suitable as an auxiliary step in combination with chemical purification	
	High temperature annealing	4. Improve crystallinity 5. High selectivity to metal	
	Chromatography, electrophoresis	6. CNTs can be separated on the bases of difference in length and conductivity	
Chemical method	Gas phase Air, Cl <sub>2</sub> , H <sub>2</sub> O, HCl, H <sub>2</sub> O, Ar, O <sub>2</sub> , C <sub>2</sub> H <sub>2</sub> F <sub>4</sub> , SF <sub>6</sub>	1. Opens the lid of the CNTs without affecting sidewalls or associated functional groups	1. Low yield 2. Produces more defects on sidewalls, breaks into different shorter length, and also the alignment and structure are affected greatly, thus limiting the final applications of CNTs
	Liquid phase HNO <sub>3</sub> , H <sub>2</sub> O <sub>2</sub> , HCl, Mixture of acid or KMnO <sub>4</sub> , Microwave in inorganic acid	2. Eliminates polyhedral and amorphous carbon and metallic impurities at the cost of substantial amounts of CNTs or damage to the CNT structure 3. Leads to functional groups	
	Electrochemical Alkali or acid solution	4. Does not disrupt or affect the alignments of CNTs	
Multi step method	Oxidation, sonication, centrifugation, filtration, wet grinding, and HIDE	1. High-purity with respect to metal 2. Metal free, improving crystallinity	1. Low yield
	Filtration/magnetic filtration, oxidation, annealing	3. Effectively removes carbonaceous and metallic impurities	
	Filtration, sonication in HNO <sub>3</sub> , HF, H <sub>2</sub> O <sub>2</sub> , or SDS	4. Better purification yield due to the early removal of metallic impurities that can oxidize CNTs	
	Annealing at high temperature, extraction		

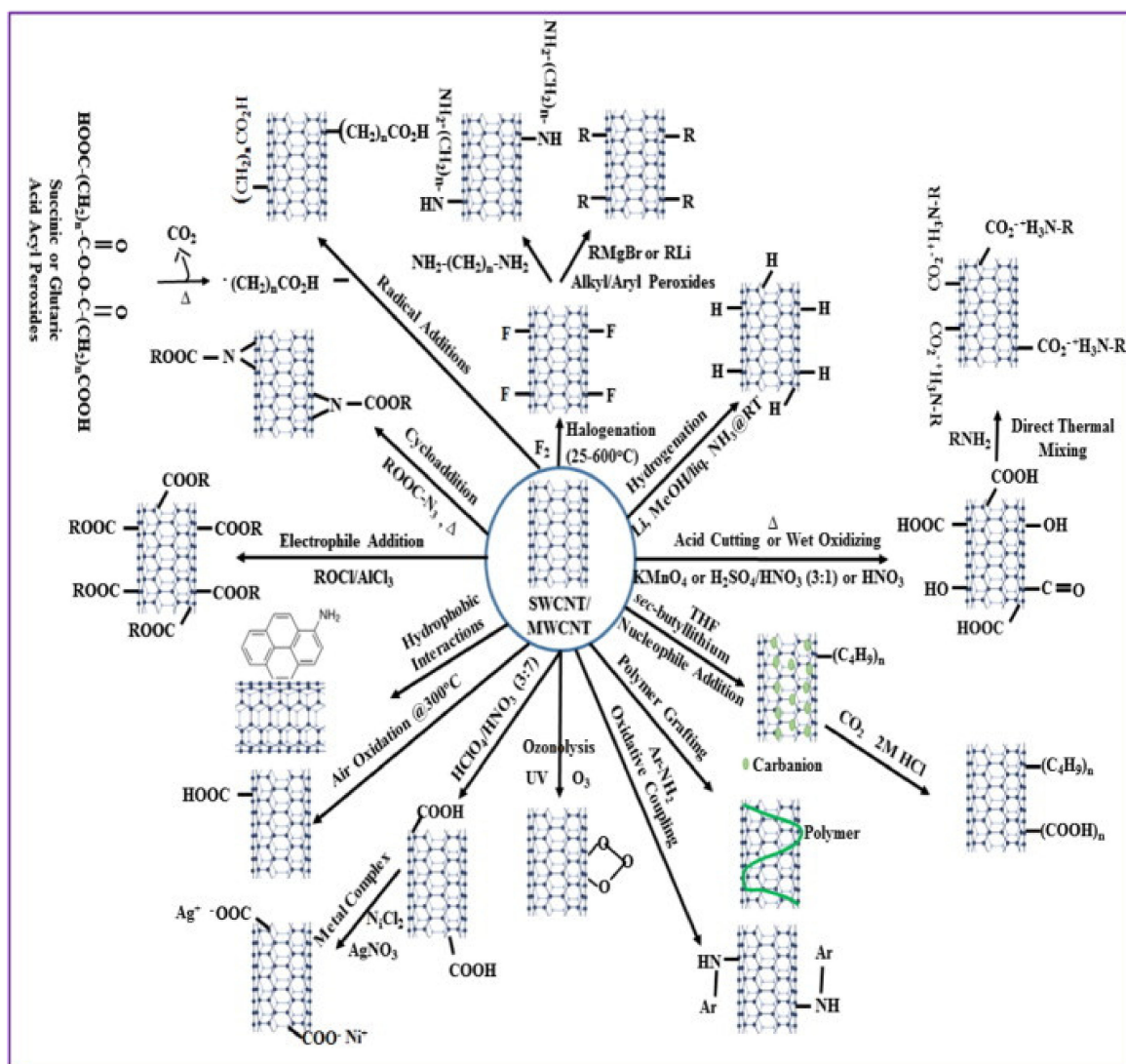
**Table 4.** Impurities associated with CNTs and their detection techniques.

Technique	Residual Material	Assessment Techniques	Advantages	Limitations
Thermo-gravimetric analysis (TGA)	Carbonaceous impurities Metal impurities	After oxidation of material, the residual metallic impurities are calculated by weighing ash and the carbonaceous impurities by area ratio of DTG	Accurate measurement of impurities	Completely oxidize/destroy the CNTs
Raman spectrometry	Carbonaceous impurities Structure defects Conductivity characteristics	The pure CNTs are associated with G-band by RBM as well as no D-band	Conductivity features and quality of CNTs can be measure along with their diameter	Difficult or even unacceptable for MWCNTs and metallic contents
Electron microscopy (SEM, TEM)	Defects in CNTs Amorphous carbon	Directly observes and qualitatively evaluates the adhesion defects on the CNT wall, the amount of amorphous carbon, fullerene	Absolute scrutiny can be undertaken	Can analyze the sample in a very small amount
UV-vis-NIR	Carbonaceous impurities conductivity characteristics	Absorption spectroscopy or reflectance spectroscopy in the ultraviolet-visible spectral region	Conductivity features and contents of CNTs can be analyzed exactly	A standard sample is needed with 100% purity
X-ray photoelectron spectroscopy (XPS)	Support material/functional groups (fine alumina, magnesium oxide, silica, zeolite, etc.)	Quantitatively characterizes the type and contents of functional groups or support materials	Analysis of functional groups on CNTs can be undertaken exactly	Unacceptable for purity
Energy-dispersive X-ray spectroscopy (EDS)	Metal impurities	Analytical technique used for chemical and/or elemental analysis of a sample	Contents and traces of different elements can be analyzed	Evaluation of the contents of CNTs is invalid

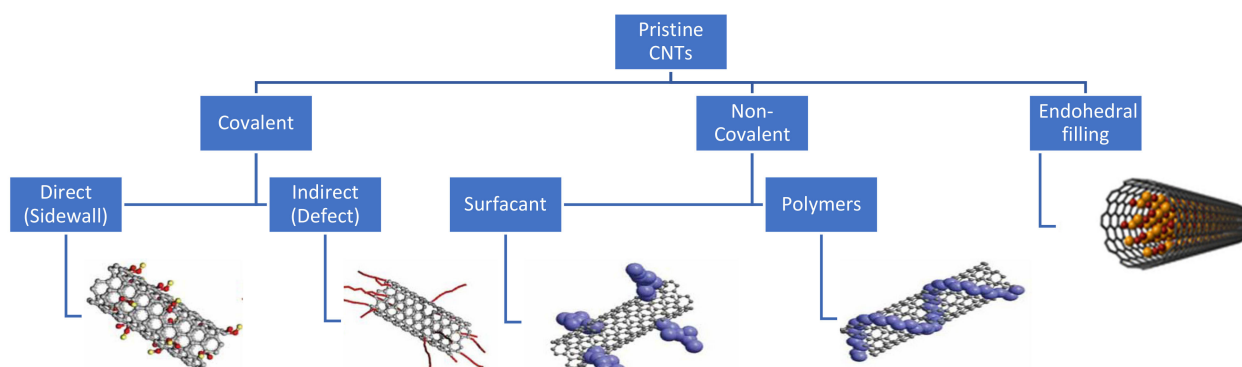
#### 4. Functionalization of CNTs

The non-polar nature of graphene layers makes the CNTs naturally hydrophobic. The hydrophobic property of CNTs is indispensable for the adsorption of aromatic contaminants like benzene and anthracene. A very strong complexation is formed due to  $\pi$  electrons present on the graphene layer making CNTs, between aromatic contaminants and the tube surface [132]. The surface affinity of CNTs can be modulated to a variety of contaminants in water and wastewater after the purification and surface functionalization. Higher adsorption of benzene was obtained by using CNTs as compared to activated carbon (AC) because of strong interaction between benzene rings and the surface of CNTs due to its hydrophobicity [133]. Figure 3 shows the different routes and schemes of CNT functionalization to increase their affinity for water and wastewater contaminants, which can be subsequently captured on the surface of CNTs used. Moreover, the functionalization of CNTs can be divided into three categories, shown in Figure 4 [134]:

1. With  $\pi$  conjugated network of CNTs through covalent bonds;
2. Attachment of different chemical groups via non-covalent bonds by using hydrophobicity of CNTs such as hydrogen bonds,  $\pi$ - $\pi$  interactions, or ionic bonds;
3. Inline filling (endohedral) of hollow tubes of CNTs. The two methods are more common for CNTs functionalization and variously used by the researchers.



**Figure 3.** Functionalization routes of CNTs and associated functional groups [132].



$\pi-\pi$

**Figure 4.** Methods of functionalization of carbon nanotubes.

CNTs are unique because of their distinctive properties such as adsorption capability, permeability, morphology, and physicochemical properties. There are several disadvantages that are also associated with raw CNTs, such as their low dispersion in solutions and low adsorption capacity for bulk fabrication of CNTs with organic and inorganic composites. In fact, aggregation is the main problem for low contaminant adsorption efficiencies by the original CNT samples [33,53,135–137] and also obscures the process of membrane preparation [138]. The  $\pi-\pi$  interactions and van der Waals force between CNTs are responsible for the less dispersion, which results in tight fit bundles and aggregation of CNTs [139]. As the number of graphite layers of CNTs decreases from MWCNTs to SWCNTs, the tendency to bunch increases [140]. Therefore, to overcome these drawbacks, the chemical reactivity and contaminant adsorption capacity of CNTs must be improved by increasing their dispersion rate, and this is done by functionalizing the nanotube [53,135–137,139,141]. Furthermore, solubility of CNTs can be increased by their functionalization, which causes them to repel each other [50]. Table 5 provides a comparison between adsorption capacities and the corresponding surface area of pristine and oxidized CNTs treated with different acids [142].

**Table 5.** Adsorption capacity of CNTs and corresponding surface area [142].

CNTs	Adsorption Capacity (mg/g)	Surface Area (m <sup>2</sup> /g)
Pristine	1.1	82.2
H <sub>2</sub> O <sub>2</sub> oxidized	2.6	130.0
HNO <sub>3</sub> oxidized	5.1	84.3
KMnO <sub>4</sub> oxidized	11.0	128.
NaOCl oxidized	47.4	94.9

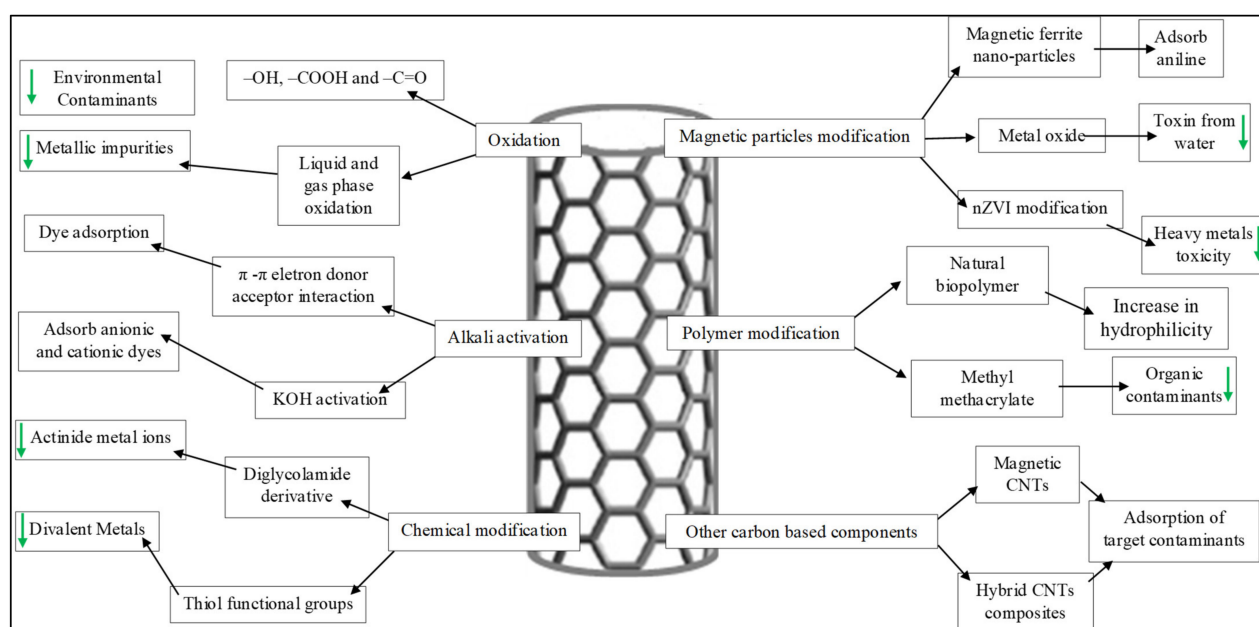
Different physical or chemical processes like oxidation, impregnation, or grafting (Figure 3) are used for the functionalization of CNTs [143,144]. During the process of functionalization, the covalent or non-covalent bonds of particular functional groups result on the end or sidewall of CNTs. Functionalization is preferred over covalent bonding, because non-covalent bonding does not influence the structure and surface area of CNTs [135]. Advantages and the limitation of covalent and non-covalent functionalization of CNTs are discussed in Table 6. Functionalization is generally done to increase the dispersibility as well as contaminant removal efficiency of CNTs, thereby improving the water or wastewater treatment application capability of the CNTs or promoting membrane fabrication [145]. Higher dispersibility in polar solvent (water) was found by covalent functionalization of CNTs with the phenolic group by 1,3 dipolar cyclo addition; covalent functionalization of CNTs with phenol groups by 1, 3-dipolar cyclo-addition was found, which facilitated

CNTs' amalgamation into the polymer matrix [139]. Oyetade et al. [141] found that MWCNTs efficiently absorbed  $Pb^{2+}$  and  $Zn^{2+}$  from aqueous samples after functionalizing with nitrogen. The increase in adsorption was due to higher surface area and more adsorption sites linked to CNTs.

**Table 6.** Benefits and limitations of covalent and non-covalent functionalization [134].

Methods	Benefit(s)	Limitation(s)
Covalent functionalization	Highly stable bonds are formed	Intrinsic characteristics are damaged Structural defects CNTs Aggregation of CNTs
Non-covalent functionalization	Simple and easy procedure CNTs structure is maintained with minimum defects Electronic characteristics of CNTs are not affected	Stability of bonds is weak

Oxidation of raw CNTs with  $HNO_3$ ,  $H_2SO_4$ ,  $HCl$ ,  $H_2O_2$ ,  $KMnO_4$ , and  $NaOCl$ , or sometimes a combination of some of these (Figure 5), has often been exercised to introduce oxidized functional groups [146,147]. Generally, oxidation improves the dispersibility and enhances the ability to adsorb certain harmful contaminants in water and wastewater at the expense of fractional damage to the surface of CNTs, as described earlier. Furthermore, the surface of CNTs can also be modified with the addition of oxygen containing functional groups by performing oxygen-plasma action. In addition to oxidation and plasma action for surface modification, the CNTs can also be successfully modified with the addition of metal oxides like  $Al_2O_3$  [148],  $MnO_2$  [149], and  $Fe_3O_4$  [150], which deliver another way to coat the surface of CNTs, thereby increasing the contaminant removal efficiencies of CNTs [53,135–137]. The potential surface modification schemes of CNTs used for targeted contaminants are shown in Figure 5.



**Figure 5.** Schematic representation of surface modification schemes of CNTs used for targeted contaminants (green arrows refer to decrease in final effluent concentration).

## 5. Characterization of CNTs

Intrinsic properties of CNTs make them fascinating and desirable candidates for diverse remediation. Characterization of CNTs plays an important role due to their distinctive properties. Numerous comments and debates have been published in past decades on

different techniques and strategies used to evaluate CNTs [151–154]. Techniques used for the characterization of CNTs are divided into four groups: microscopy and diffraction, thermal, spectroscopic, and separation techniques [155]. Sometimes research includes one more group, the magnetic measurement technique. Table 7 shows different characterization techniques used for the evaluation of CNTs. It must be noted more than one technique is prominent for the characterization of CNTs; techniques employed alone are not fully characterized, nor they are absolutely quantitative. Even though qualitative analysis of CNTs can be done by electron microscopy, scanning electron microscopy (SEM) evaluates the nanostructure of the tubes, and transmission electron microscopy (TEM) is used for further precise inspection, generally detecting the defects in CNTs [129,156].

**Table 7.** Different analytical techniques used for the characterization of carbon nanotubes [155].

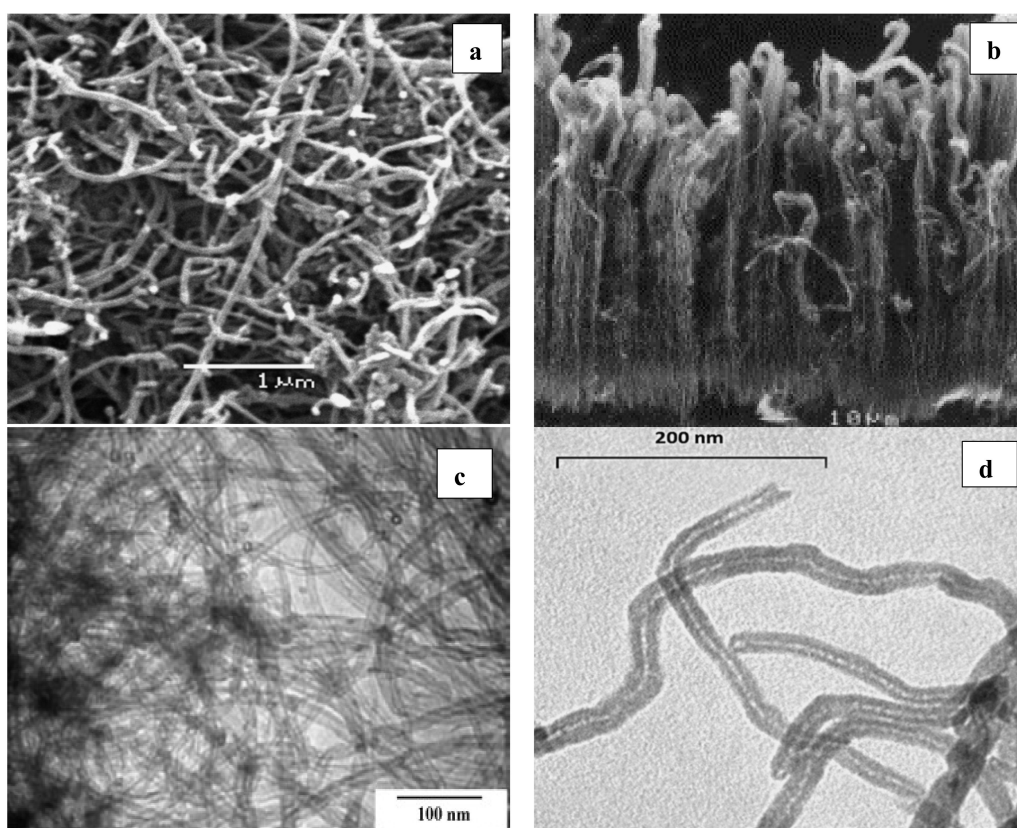
Characterization Techniques	Used for Studying
Microscopy and diffraction techniques	[157–159]
SEM	Morphological analysis (diameter and length), aggregation state
TEM/HR-TEM	Morphological analysis of internal structure (diameter, number of layers and distance between them)
AFM	Morphological analysis of internal structure (diameter, number of layers and distance between them)
Scanning tunneling microscopy	Morphological analysis of internal structure (diameter, number of layers and distance between them)
Neutron diffraction	Morphological analysis of bulk samples
XRD	Morphological analysis of bulk samples
Spectroscopic techniques	[139,160,161]
Raman spectroscopy	Purity and presence of by-products, diameter distribution, (n, m) chirality
IR and FT-IR	Purity, functionalization by attaching functional groups to the sidewall
UV–vis and NIR	Dispersion efficiency, diameter and length distribution, purity
Fluorescence spectroscopy	Size, dispersion efficiency, (n, m) chirality
XPS and EDS	Elemental composition, functionalization (covalent and non-covalent)
Thermal techniques	[162]
TGA	Purity and presence of by-products, quality control of synthesis and manufacture processes
Separation techniques	
Size exclusion chromatography	Purification, separation by size (length)
Capillary electrophoresis	Purification, separation by size (length, diameter, and cross-section)
Field flow fractionation	Fractionation by size (length)
Ultracentrifugation	Separation by chirality, electronic type, length, and enantiomeric identity
Magnetic techniques	[158,163,164]
Vibrating sample magnetometry	Magnetic properties
Alternating gradient magnetometry	Magnetic properties
Superconducting quantum interference device	Magnetic properties

The ultrastructure of different types of species including organic, inorganic, and biological species can be evaluated by using very popular techniques known as SEM and TEM. The scanning of the sample in SEM generates an image when the targeted electron ray interacts with sample of CNTs. Generally, the technique is used to analysis the morphology (length and diameter) of nanomaterials [123,155,165] to assess the quality of prepared CNTs. For example, Figure 6a shows a SEM image of the as-prepared CNTs [148], and Figure 6b



shows alignment of CNTs synthesized by using a horizontal quartz tube housed in muffle furnace. Average length of the tubes was 70  $\mu\text{m}$  measured by using SEM [166]. Sometimes, it is also used to validate the surface modifications in terms of functional reactions that occur on the surface of CNTs [167]. In the case where the required measurement exceeds 1 to 20 nm resolution while using SEM, then TEM is used [156,167]. Small dimensions in CNTs such as interlayer distance, diameter, and number of graphene sheets can be easily examined (Figure 6c,d) by targeting the sample with a high energy electron beam of up to 300 keV [168,169]. It should be noted that the functional groups (organic and inorganic) that modify the surface of CNTs can also be evaluated by using TEM [161,170,171]. Moreover, structural integrity, surface functionalization, and defect son the surface of CNTs caused by the oxidation (acidic or basic) to introduce oxygen containing functional groups like hydroxyl, carbonyl, and carboxylic acid groups have also been studied by using both SEM and TEM techniques [155,172].

The image of the atomic structure and crystal structure information can be obtained by using high resolution transmission electron microscopy (HR-TEM) [173]. A high phase contrast image as small as the crystal unit can be obtained. This technology is widely used for advanced characterization of materials, allowing access to information on just-in-time defects, stacking faults, deposits, and grain boundaries. In addition to the morphology of the MWCNTs in the HR-TEM image, the direct measurements can be made on the MWCNT walls. For example, the number of walls constituting the nanotubes can be directly counted and recorded as control parameters in subsequent experiments in case the number of such walls needs to be changed. In addition, the interplanar distance between the walls can be accurately measured and compared with the crystal structure data table of the carbon structure and its respective diffraction pattern.



**Figure 6.** (a) SEM image of a bulk sample of multi-walled carbon nanotubes [155], (b) SEM image showing vertical aligned CNTs [174], (c) high-magnified TEM images of CNTs grown on unreduced catalyst [168] and, (d) TEM image of a bulk sample of multi-walled carbon nanotubes [155].

The chemical state and structure of CNTs can be obtained by using an X-ray photoelectron spectroscopy (XPS) technique [114]. However, the data obtained from this technique are used to examine the structural modification of CNTs before and after the chemical interactions with organics, inorganics, or gaseous adsorption. The investigation of CNTs by using XPS is done after the incorporation of nitrogen into the nanotube [175]. Due to the polar nature of the carbon–nitrogen bond, the peak shift before and after the modification is evidence of nitrogen incorporation [176,177]. Furthermore, the technology demonstrates that carbon nanofibers are more similar to carbon oxides than various graphites [178]. Fluorinated functionalization of SWCNTs was also studied by using XPS; the results concluded that three peaks of  $sp^2$ ,  $sp^3$ , and carboxyl groups (284.3, 285, and 288.5 eV, respectively) were associated with C1s of un-doped nanotubes [114]. The observed carboxyl group (288.5 eV) was similar to in nanofibers [179].

Crystallinity of CNTs can be obtained by using an X-ray diffraction (XRD) technique [90,176]. X-ray diffraction patterns of graphite and CNTs are very close to each other because of their inherent properties. The XRD pattern of CVD-synthesized MWCNTs is shown in Figure 7, illustrating a peak similar to graphite (002), and the measured layer spacing can be obtained from Bragg's law, while the other peaks (family of (hk 0) peaks)) can be obtained because the mono-graphene layer makes the honeycomb matrix [114]. Therefore, the curve obtained by XRD does not distinguish between the microstructure information of CNTs and graphite; nevertheless, it is helpful for purity analysis of the sample [102,103]. The XRD pattern revealed that straight CNTs with a good alignment on the surface of the substrate did not show the peak, i.e., 022 [114]. For CNTs aligned vertically to their substrate surface, the XRD pattern is not collected because of the scattering of the beams inside the sample. Therefore, the 002 peak is always lowered for better-aligned CNTs [180]. In addition, several other types of parameters such as bundle size, mean diameter, and diameter dispersion can also be studied by using the XRD technique. The peak 10 is greatly influenced by these parameters in terms of its location and thickness [181,182].

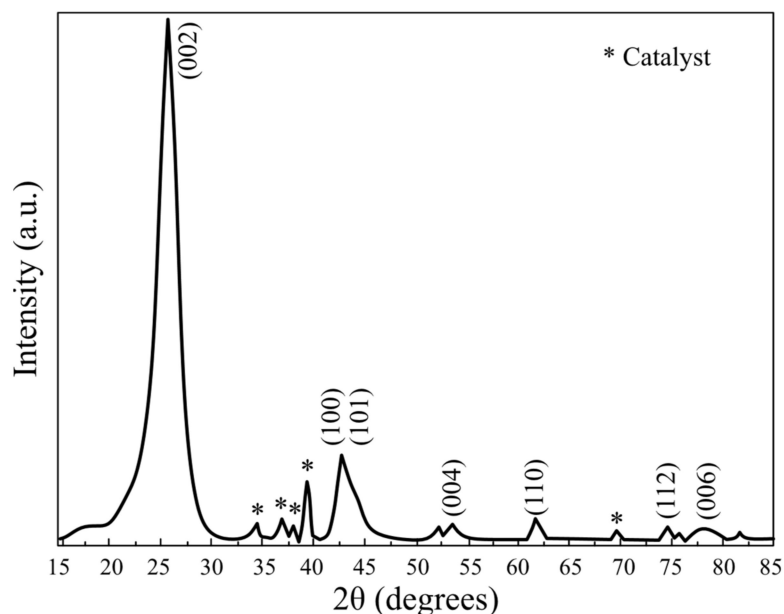
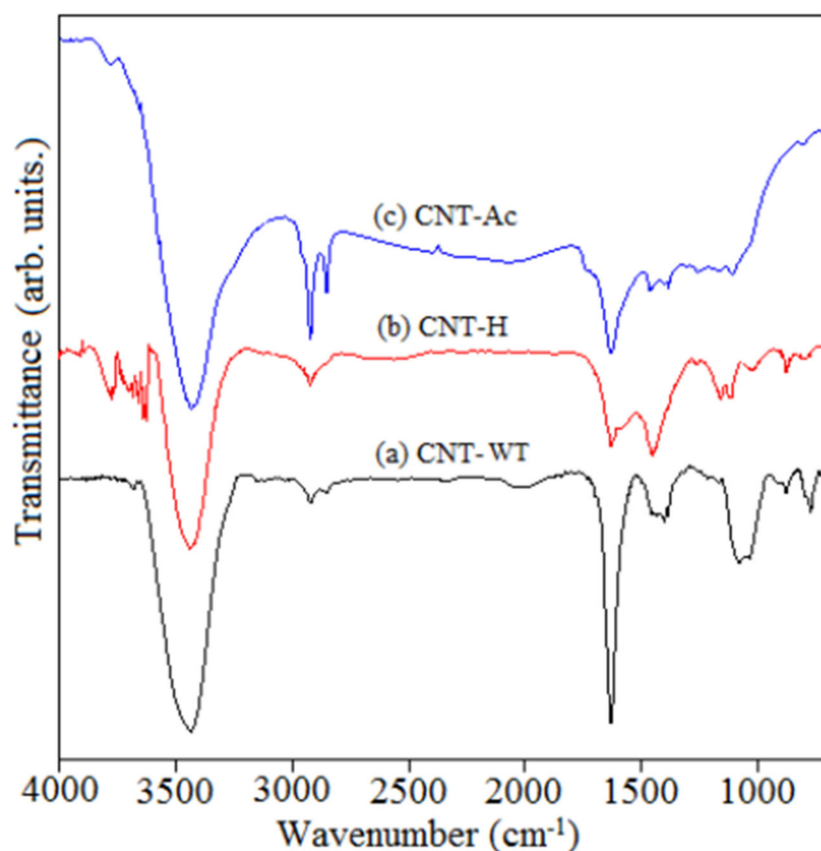


Figure 7. XRD pattern of MWCNTs synthesized by CVD [114].

Qualitative evaluation for nanotubes' surface can be done by using Fourier transform infrared spectroscopy (FT-IR) [183]. The sample is characterized by passing the infrared radiations through it, and the part of radiations absorbed by the sample at specific energy is determined. Each functional group is identified by a particular range of frequency with associated absorption peaks. The infrared spectrum of the original CNTs (Figure 8) shows a characteristic band of about  $1600\text{ cm}^{-1}$  associated with aromatic rings (C=C bond) of

rolled graphene layers. Sometimes, the peaks of 3800 to 3200  $\text{cm}^{-1}$  (O-H stretch) and 1700  $\text{cm}^{-1}$  that absorb moisture into the atmosphere or due to certain purification processes [184]. In addition, a band of 2910 to 2940  $\text{cm}^{-1}$  was also detected for CNTs, which is related to the vibration of C-H stretching methylene ( $\text{CH}_2$ ) [185]. The FTIR spectra can also suggest surface modification of CNTs. For example, in Figure 8b, a new peak near 1450  $\text{cm}^{-1}$  appeared that was assigned to asymmetric  $\text{CH}_2$  bending. This peak is typically interpreted as evidence of defects in the structure of CNTs. In Figure 8c, a new peak near 1735  $\text{cm}^{-1}$  suggested a carbonyl stretch of the carboxylic acid group. In addition, a double peak at approximately 2900  $\text{cm}^{-1}$  was attributed to the loss of aromaticity due to the oxygen functional groups. The thermal stability and proportion of volatile compounds of nanotubes can be analyzed by an analytical technique called thermogravimetric analysis (TGA). The analysis is made by heating the sample directly in the air or inert gases (He/Ar) while recording the change in its weight with respect to elevated temperature [176,186,187]. In some cases, the analysis is made in the presence of  $\text{N}_2$  or He with poor oxygen atmosphere (1% to 5%  $\text{O}_2$ ) to slow oxidation [188]. During the TGA analysis of the CNT sample in the air atmosphere, the weight loss of the sample (Figure 9) is usually due to carbon oxidation to  $\text{CO}_2$ , while the solid oxides after the oxidation of metallic catalyst are responsible for the superposition of the sample [165,189,190]. Generally, the percent yields of carbon deposits are determined by using TGA. Usually, during the oxidation of the sample the weight occurs in the temperatures ranging from 200 to 680  $^\circ\text{C}$  [190]. The contents of carbon can be calculated by obtaining the percentage of  $(m_1 - m_2)/m_1$ , where  $m_1$  is the weight of the sample before oxidation and  $m_2$  is the weight of the sample after the oxidation [11].



**Figure 8.** FT-IR spectra of CNTs: (a) pristine CNTs (CNTs-WT), (b) CNTs treated with HCl (CNT-H), and (c) CNTs treated with mixture of  $\text{H}_3\text{SO}_4$  and  $\text{HNO}_3$  (CNTs-AC) [191].

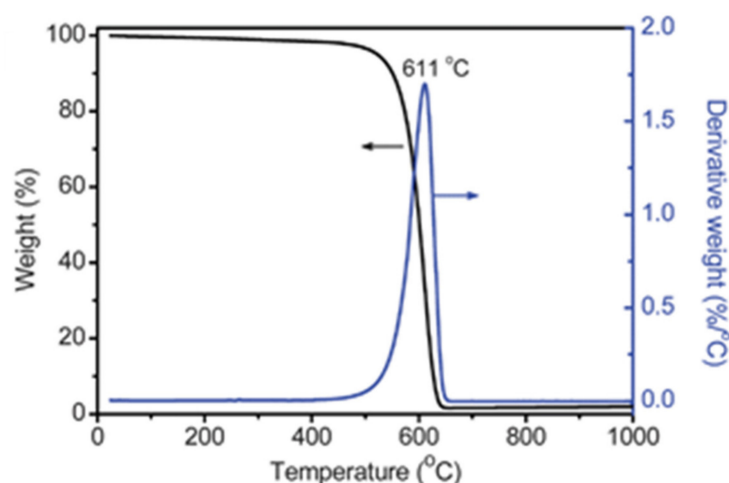


Figure 9. TGA analysis of the CNT sample [190].

The structural features of CNTs can be defined by their crystalline structure, chiral carrier, and single- or multi-walled features [192]. Crystalline arrangement of nanotubes can be characterized by the ID/IG ratio determined from the Raman spectroscopy. ID/IG indicates the ratio between the organized and unstructured carbon in the CNTs and uses the intensity of defective and graphitic carbon (D and G bands) at the high wavenumbers in the Raman absorption spectrum (Figure 10) [193]. ID/IG is a good quality indicator for CNTs, and the low ID/IG ratio is characteristic of highly graphitized structures; the laboratory reported mass of MWCNT is 0.65, and industrial grade MWCNT is 2.04 [90,194].

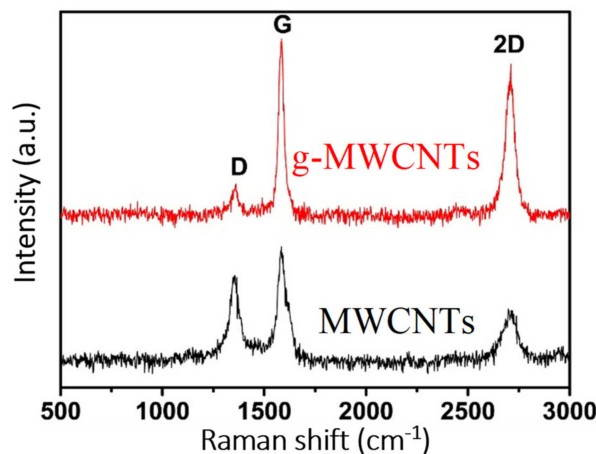
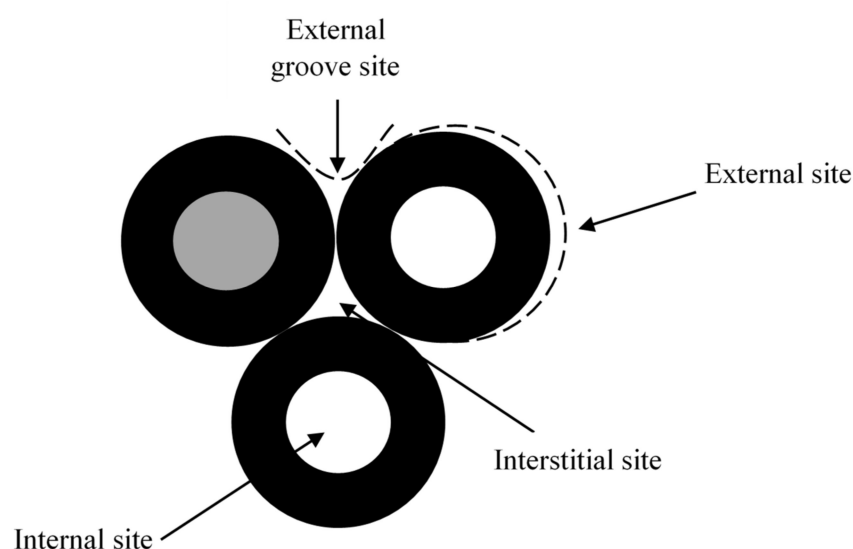


Figure 10. Raman spectra of pristine (MWCNTs) and graphitized MWCNTs (g-MWCNTs) [193].

## 6. Applications of CNTs

### 6.1. Removal of Heavy Metals

Removal of heavy metals from water and wastewater by using surface modified CNTs has been studied extensively [143,195–197]. Contaminant adsorption mainly occurs at four possible types of sites that are available on the CNTs such as outer and inner grooves and interstitial channels as shown in Figure 11, but the inner region of the nanotubes is less adsorbed [53,135–137,198]. Bahgat et al. [143] used functionalized MWCNTs for the adsorption of heavy metals and concluded that adsorption of the metals occurs because of a number of adsorption spots available on the tubes' surfaces. In another study, higher adsorption of  $\text{Zn}^{2+}$  ions was observed by the plasma-treated CNTs due to the availability of more oxygen-containing functional groups. The mechanism of surface complexation was responsible for the adsorption of cationic ions, as the adsorption sites increased due to deprotonating functional groups [197].



**Figure 11.** Different adsorption sites on CNTs [198].

Surface modification of CNTs improved their affinity for heavy metal ions and metalloid adsorption [199,200] by incorporating iron oxide and oxidation [16,201–203], coating with oxides of nonmagnetic metals [204], introducing a thiol functional group [82], and functionalizing with sulfur [205]. According to Addo Ntim and Mitra [203], the different oxides of iron such as magnetite ( $\text{Fe}_3\text{O}_4$ ), maghemite ( $\gamma\text{-Fe}_2\text{O}_3$ ), hematite ( $\alpha\text{-Fe}_2\text{O}_3$ ), and goethite ( $\alpha\text{-FeO}$ ) are very useful for the removal of trace heavy metals present in water and wastewater [170].

Generally, the process of surface modification or functionalization of CNTs is responsible for the adsorption mechanisms of heavy metals [53,191,206]. Table 8 shows an explanation of previously used functionalized CNTs and their adsorption capacities and removal mechanisms. Surface area, porous structure, functional groups, and interaction mechanisms between the adsorbate and adsorbents are the main attributes of CNTs for their heavy metal adsorption capability [53,136,137,143]. In addition, these properties enhance the heavy metal removal efficiencies onto the polymer film incorporated by functional CNTs [135].

Physical and chemisorption mechanisms of adsorption have been observed between the CNTs modified with metal oxides and heavy metal ions [203]. In addition, the contaminant with particular hydration energy, ionic radius, and potential of hydrolysis also affects the adsorption capacities of CNTs. A study conducted by Hu et al. [207] confirms the above statement, when higher removal of  $\text{Pb}^{+2}$  was observed than  $\text{Cu}^{+2}$  by using MWCNTs modified with iron oxide under the same experimental conditions. The adsorption performance of CNTs for heavy metals is also affected by the presence of other organics. For example, enhanced removal of  $\text{Cd}^{+2}$  ion was observed by using oxidized MWCNTs in the presence of 1-naphthol, while there was no effect recorded during the presence of  $\text{Cd}^{+2}$  ion on the removal of 1-naphthol in similar conditions [208]. These types of results mainly occur because of the distinct interaction mechanisms that are responsible for the adsorption of different types of pollutants [170].

Solution pH also plays an important role in the adsorption capacity and mechanisms of CNTs. For example, higher pH is favorable for endospheric interactions, while the lower pH facilitates extracellular interaction and/or ion exchange of targeted metal ions and surface functional groups of CNTs [208]. Moreover, as the pH increases (alkaline conditions), the charge density of functionalized CNTs moves towards more negativity, which is efficient for the adsorption of cationics, while, at lower pH (acidic conditions), because of protonating functional groups, the positive charge density increases, which repels the cationic metals, resulting in the lower efficiency of CNTs [53,136,137]. In a study, oxidized and ethylenediamine-doped MWCNTs were used for the adsorption of



$\text{Cd}^{+2}$  ions from aqueous samples. The results concluded that both types of absorbents removed  $\text{Cd}^{+2}$  ions strongly depending on pH ranging from 8 to 9 [209]. According to Rao and coworkers [146], the best adsorption capacity of nanotubes was observed in the pH ranging from 7 to 10. In addition to this particular range of pH, the ionization and competition between different species may also happen [210–213]. For example, an effective adsorption of  $\text{Pb}^{+2}$  ions on functionalized CNTs was reported during the co-existence of sodium dodecylbenzenesulfonate (DBS), while the adsorption of lead was significantly reduced in the presence of benzalkonium chloride [214]. The higher adsorption of lead ions might be due to anionic surfactant formation between  $\text{Pb}^{+2}$  and DBS, which are very complex compounds. In addition, the charge density (negative or positive) on the surface of CNTs makes the different interaction mechanisms of metal ions. For example, chemical interaction occurred between the N-doped magnetic CNTs and Cr(III), while an electrostatic interaction was observed between the acid oxidized CNTs and Cr(III) [215].

Regardless of the costs, the CNTs are more efficient in terms of their adsorption and desorption phases than the AC, as reported in many studies [146,216]. Lu and coworkers [216] reported a slight decrease in adsorption and desorption of  $\text{Ni}^{+2}$  while using CNTs, and on the other hand, a sharp decrease was observed in the case of AC [216].

The adsorption of heavy metals by the functional CNTs depends on the interactions between available functional groups on the surface of CNTs and particular contaminants rather than the size of nanotubes. For example, higher adsorption of As(III) and As(V) was reported by Addo Ntim and Mitra [203] by using zirconia-modified CNTs of the same diameter (20 to 40 nm) than by using the CNTs coated with iron oxide. These results demonstrate higher arsenic adsorption by zirconia nanocomposites than the CNTs modified with iron oxide with similar range of diameters. Therefore, based on these results, further investigations on the adsorption capacities of functionalized CNTs are needed in terms of their surface area rather than the size CNTs.



**Table 8.** Functional CNT based nanomaterials used for the removal of heavy metal ions from aqueous samples.

Adsorbate	Adsorbent	Surface Area (m <sup>2</sup> /g)	Diameter (nm)	Q <sub>e</sub> /RE	Experimental Conditions			Removal Mechanism	Model	Comments	Ref.
					pH	IC	AL				
As(III)	MWCNTs	9.1	10–40	91%	6	40 µg/L	2.0 g/L	Liquid film diffusion, ion exchange	Tempkin, Dubinin-Radushkevich, Langmuir, Freundlich	In column operation, the removal As(III) was up to 13.5 µg/L	[217]
	Zero-valent iron-doped MWCNT	-	-	200 mg/g	4	10 mg/L	0.2–4.0 g/L	Ion exchange, surface complexation	Langmuir	Maximum As(III) removal efficiency was 98.5%	[218]
	Floating catalyst CNTs (FCNT)	74	18.6	1.22 mg/g	6.5	0.1–10 mg/L	1 g/L	Electrostatic attraction, surface complexation	Langmuir	Potential adsorbent for removal to total arsenic	[219]
	Oxidized-FCNT	129	10.7	1.90 mg/g							
	Heat-treated oxidized CNTs (FCNT-HOX)	168	7	5.99 mg/g	7	0.1–1 mg/L	0.05 g	Surface complexation	Langmuir	Reusability of adsorbent was up to 5 cycles	[219]
	Zero-valent iron immobilized on MWCNTs	78.78		111.1 mg/g							
	MWCNT-ZrO <sub>2</sub>	152	20–40	2 mg/g	6	100 µg/L	100 mg/10 mL	Chemisorption/ physisorption	Langmuir	The adsorption capacity of AS (III) is not associated with pH value	[203]
	Iron-oxide-coated multi-walled carbon nanotubes	153	20–40	1.723 mg/g	4	100 µg/L	10 mg/10 mL	Electrostatic interaction, surface complexation	Langmuir	Suggesting that modifying MWCNTs with other groups can develop potential adsorbents for water treatment	[203]

Table 8. Cont.

Adsorbate	Adsorbent	Surface Area (m <sup>2</sup> /g)	Diameter (nm)	Q <sub>e</sub> /RE	Experimental Conditions			Removal Mechanism	Model	Comments	Ref.
					pH	IC	AL				
As(V)	Iron-oxide-coated SWCNTs	-	-	49.65 mg/g	4	5–50 mg/L	-	Surface complexation	Freundlich	Adsorption was very fast for low concentration of As(V)	[220]
	MWCNTs	9.1	10–40	92%	6	40 µg/L	2.0 g/L	Liquid film diffusion, ion exchange	Tempkin, Dubinin-Radushkevich, Langmuir, Freundlich	In column operation, the removal As(III) was up to 14.0 µg/L	[217]
	Zero valent iron doped MWCNTs	-	-	250 mg/g	4	10 mg/L	0.2–4.0 g/L	Ion exchange, surface complexation	Langmuir	Maximum As(V) removal efficiency was 98.5%	[218]
	Floating catalyst CNTs (FCNT)	74	18.6	0.88 mg/g	6.5	0.1–10 mg/L	1 g/L	Electrostatic attraction, surface complexation	Langmuir	Potential adsorbent for removal to total Arsenic	[219]
	Oxidized-FCNT	129	10.7	2.51 mg/g							
	Heat-treated oxidized CNTs (FCNT-HOX)	168	7	6.37 mg/g							
	3-(2-aminoethylamino) propyltrimethoxysilane modified MWCNTs	-	-	8.01 mg/g	2.2	1.0 mg/L	40 mg	-	-	Cr(IV) was selectively adsorbed in the micro-column packed with adsorbent	[221]
	Zero-valent iron immobilized on MWCNTs	78.78		167 mg/g	7	0.1–1 mg/L	0.05 g	Surface complexation	Langmuir	Successfully applied to ground water with high pH	[219]

Table 8. Cont.

Adsorbate	Adsorbent	Surface Area (m <sup>2</sup> /g)	Diameter (nm)	Q <sub>e</sub> /RE	Experimental Conditions			Removal Mechanism	Model	Comments	Ref.
					pH	IC	AL				
	Iron(III)-oxide-coated ethylenediamine functionalized MWCNTs	198.5	5–10	23.5 mg/g	4	100 µg/L	50 mg/10 mL	Ion exchange	Langmuir	Greater efficiency to remove As(V) due to enormous adsorbing sites	[222]
	MWCNT–zirconia nanohybrid	152	20–40	5.0 mg/g	6	100 µg/L	100 mg/10 mL	Chemisorption/ physisorption	Langmuir	The adsorption capacity of As(V) is not associated with pH value	[223]
	Iron-oxide-coated MWCNTs	153	20–40	0.189 mg/g	4	100 µg/L	10 mg/10 mL	Electrostatic interaction, surface complexation	Langmuir	Modifying MWCNTs with other groups can develop potential adsorbents for water treatment	[203]
	Iron oxide/carbon nanotubes/chitosan magnetic composite film	64.4		66.25 mg/g	2–10	100 mg/L	0.3 mg/g	Electrostatic	Langmuir	Decrease in efficiency was 12% after reusing the adsorbent for ten cycles	[224]
	Cr(III) Nitrogen-doped magnetic carbon nanoparticles	-	-	83.7 mmol/g	8	200 mg/L	10 mg/500 mL	Chemical adsorption	Langmuir	10-fold greater removals than activated carbon due to large SSA	[225]
	Acid modified MWCNTs	-	23	0.5 mg/g	7	1 mg/L	120 mg/500 mL	Electrostatic interaction	Pseudo-second order	Increasing removal of Cr with increasing the dose of CNTs	[215]

Table 8. Cont.

Adsorbate	Adsorbent	Surface Area (m <sup>2</sup> /g)	Diameter (nm)	Q <sub>e</sub> /RE	Experimental Conditions			Removal Mechanism	Model	Comments	Ref.
					pH	IC	AL				
Cr(IV)	Iron oxide/carbon nanotubes/chitosan magnetic composite film	64.4	-	449.3 mg/g	10-Feb	100 mg/L	0.3 mg/g	Electrostatic	Langmuir	Decrease in efficiency was 6% after reusing the adsorbent for ten cycles	[224]
	Nitrogen-doped magnetic CNTs	116.4	-	970.9 mg/g	1	40–1000 mg/L	0.5–3.5 g/L	Surface complexation	Langmuir	Recycled adsorbent was successfully used for excellent electrochemical reduction of CO <sub>2</sub>	[163]
	Chitosan-modified MWCNTs	-	30–50	164.0 mg/g	2	50 mg/L	50 mg	Electrostatic	Langmuir	Adsorbent can be recycled up to 4 times	[226]
	Magnetic iron oxide MWCNTs	-	~50	42.0 mg/g	2	5 mg/L	0.4–1.0 g/L	Electrostatic	Langmuir	Absorbent highly showed durability, selectivity, easy regeneration ability	[227]
	Chitin magnetite MWCNTs	69.1	-	100%	2	50 mg/L	-	Physical	-	Removal of Cr(IV) was enhanced after mixing MWCNTs with chitin	[228]

Table 8. Cont.

Adsorbate	Adsorbent	Surface Area (m <sup>2</sup> /g)	Diameter (nm)	Q <sub>e</sub> /RE	Experimental Conditions			Removal Mechanism	Model	Comments	Ref.
					pH	IC	AL				
	Magnetic MWCNTs	200	20–40	16.23 mg/g	3	25 mg/L	-	-	Langmuir	The adsorption capacity of adsorbent increases with initial concentration of Cr(VI) and contact time, but decreases with the increase of adsorbent dosage	[229]
	3-(2-aminoethylamino) propyltrimethoxysilane-modified MWCNTs	-	-	9.79 mg/g	2.2	1.0 mg/L	40 mg	-	-	Cr(IV) was selectively adsorbed in the micro-column packed with adsorbent	[221]
	Activated-carbon-coated CNTs	-	10–20	9.0 mg/g	2	0.2–0.5 mg/L	2 mg/50 mL	-	Langmuir	The f-CNT can be used largely for the removal of Cr ions	[215]
	Ceria-supported CNTs nanoparticles	-	20–80	31.55 mg/g	7	35.3 mg/L	100 mg/100 mL	Ion exchange	Langmuir	Suggesting that CeO <sub>2</sub> /ACNTs has high potential for heavy metal removals	[230]

Table 8. Cont.

Adsorbate	Adsorbent	Surface Area (m <sup>2</sup> /g)	Diameter (nm)	Q <sub>e</sub> /RE	Experimental Conditions			Removal Mechanism	Model	Comments	Ref.
					pH	IC	AL				
Pb(II)	Thiol-functionalizedM-WCNTs/Fe <sub>3</sub> O <sub>4</sub>	97.367	-	65.4 mg/g	6.5	50 mg/L	100 mg/100 mL	Lewis acid–base interactions	Langmuir	The adsorbent removed heavy metal ions effectively at various pH values	[231]
	Magnetic MWCNTs	295.4	-	N/A	6	100 mg/L	1000 mg	-	Experimental	High removal efficiency due to intrinsic properties, large SSA, and porous structure	[232]
	MWCNTs/Fe <sub>3</sub> O <sub>4</sub>	108.37	10–20	22.04 mg/g	5.3	30 mg/L	500 mg/1000 mL	Electrostatic, hydrophobic, and $\pi$ – $\pi$ interactions	Langmuir	Easily regenerate the adsorbent by external magnetic field after several cycles	[233]
	MWCNTs/Fe <sub>3</sub> O <sub>4</sub> modified with 3-aminopropyltriethoxysilane	90.68	10–20	75.02 mg/g	5.3	30 mg/L	500 mg/1000 mL	Electrostatic, hydrophobic, and $\pi$ – $\pi$ interactions	Langmuir	Easily regenerate the adsorbent by external magnetic field after several cycles	[233]
	MWCNTs grafted/PAAM membrane	-	-	98%	-	10 mg/L	1000 mg/1000 mL	Electrostatic interaction	-	The f-CNT membrane potentially enhances the water flux and removal of heavy metals	[234]



Table 8. Cont.

Adsorbate	Adsorbent	Surface Area (m <sup>2</sup> /g)	Diameter (nm)	Q <sub>e</sub> /RE	Experimental Conditions			Removal Mechanism	Model	Comments	Ref.
					pH	IC	AL				
	Oxidized CNT sheets	-	-	117.65 mg/g	7	1200 mg/L	50 mg/25 mL	Chemical interaction	Langmuir	Considering the oxidize CNT sheets promising nanomaterial for adsorption	[235]
	MWCNTs grafted with 2-Vinylpyridine	-	-	37.0 mg/g	6	10 mg/L	640 mg/1000 mL	Ion exchange, electrostatic interaction	Langmuir	Showed high suitability for preconcentration and immobilization of heavy metal ions from water	[236]
	Oxidized MWCNTs	142.29	10–30	0.021 mmol/g	4.1	0.83 mmol/L	0.75 g/L	Chemical, electrostatic, hydrophobic, and $\pi$ - $\pi$ interactions	Langmuir	High removal efficiency toward heavy metal ions in wastewater	[214]
	Alumina-coated MWCNTs	-	-	99%	Different	-	10 mg/25 mL	N/A	-	The composite can be used largely to remove lead from industrial wastewater. Adsorption efficiency increased with the pH (3 to 7)	[236]
	Nitrogen-doped magnetic carbon nanoparticles		-	6.74 mmol/g	8	200 mg/L	10 mg/500 mL	Chemical adsorption	Experimental	High removal efficiency toward Pb compared to Cr	[225]

Table 8. Cont.

Adsorbate	Adsorbent	Surface Area (m <sup>2</sup> /g)	Diameter (nm)	Q <sub>e</sub> /RE	Experimental Conditions			Removal Mechanism	Model	Comments	Ref.
					pH	IC	AL				
	Titanium Dioxide /MWCNT composites	-	-	137.0 mg/g	6	10 mg/L	20 mg/10 mL	-	Langmuir	Important adsorption ability to remove large amount of Pb(II) in short period	[204]
Pb(II)	Oxidize MWCNTs	-	20–30	-	-	10 mg/L	3000 mg/1000 mL	-	-	The sorption of Pb largely depends on foreign ions and ionic strength	[237]
	Manganese oxide-coated CNTs	275	2.60	78.74 mg/g	5	30 mg/L	50 mg/100 mL	Electrostatic interaction, surface complexation	Langmuir	300% greater adsorption capacity than raw CNTs	[238]
	Acidified MWCNTs	237.3	29.0	85 mg/g	5	50 mg/L	25 mg/50 mL	Physical adsorption	Langmuir	The regeneration of Pb increasing with decreasing pH and can be used for several cycles	[239]
Cd(II)	Alumina-decorated MWCNTs	109.8	10–20	27.21 mg/g	7	1 mg/L	50 mg/L	Electrostatic interaction, physical adsorption, surface precipitation	Langmuir	Capable of removing both metallic and organic Contaminants	[240]

Table 8. Cont.

Adsorbate	Adsorbent	Surface Area (m <sup>2</sup> /g)	Diameter (nm)	Q <sub>e</sub> /RE	Experimental Conditions			Removal Mechanism	Model	Comments	Ref.
					pH	IC	AL				
	Oxidized MWCNTs	78.5	16.09	24.15 mg/g	-	5 mg/L	1 mg/10 mL	Chemisorption	Langmuir	The sorption capacity is strongly dependent on pH due to surface charge and showed best performance in the pH ranging from 6 to 10	[209]
	Ethylenediamine-functionalized MWCNTs	101.2	21.25	25.7 mg/g	-	5 mg/L	1 mg/10 mL	Chemisorption	Langmuir	The sorption capacity is strongly dependent on pH due to surface charge and showed best performance in the pH ranging from 6 to 10	[209]
	Oxidized CNT sheets	-	-	92.59 mg/g	7	1200 mg/L	50 mg/25 mL	Chemical interaction	Langmuir	Excellent removal of heavy metal ions	[235]
	Acid-modified CNTs	170	10–20	4.35 mg/g	7	-	50 mg	Electrostatic interaction	Langmuir	Potential material for water purification	[241]
	MWCNTs modified with Chitosan	-	60–100	-	-	-	2000 mg	Electrostatic interaction	-	The removal efficiency increases with increase of mass of both MWCNTs and chitosan	[242]

Table 8. Cont.

Adsorbate	Adsorbent	Surface Area (m <sup>2</sup> /g)	Diameter (nm)	Q <sub>e</sub> /RE	Experimental Conditions			Removal Mechanism	Model	Comments	Ref.
					pH	IC	AL				
Hg(II)	MnO <sub>2</sub> -coated CNTs	110.4	30–50	58.82 mg/g	5–7	10 mg/L	200 mg/20 mL	Electrostatic interaction	Langmuir	Higher adsorption affinity to other heavy metals rather than Hg	[149]
	Thiol-derivatized SWCNTs	-	-	131.58 mg/g	5	40 mg/L	0.25 mg/mL	Electrostatic interaction	Langmuir	Easily desorb/regenerate Hg after treatment of water	[243]
	Amino and thiolated functionalized-MWCNTs	-	5–10	84.66 mg/g	6	100 mg/L	60 mg	Physisorption	Langmuir	Highly efficient removal from real wastewater and further research is necessary to commercialize	[244]
	Iodide-incorporated MWCNT (CNT-I)	153	10–20	123.45 mg/g	6	100–500 mg/L	2500 mg/1000 mL	Ion exchange	Langmuir	Successfully used for the adsorption and desorption of Hg(II)	[205]
	Sulphur-containing CNTs	-	-	72.8 µg/g	12.15	0.1mg/L	100 mg/20 mL	Chemisorption	Freundlich	Greater treatment ability for industrial wastewater containing Hg and other anions and cations	[245]
Hg(II)	Thiol-functionalized-MWCNTs/Fe <sub>3</sub> O <sub>4</sub>	97.2	-	65.52 mg/g	6.5	50 mg/L	1000 mg/100 mL	Lewis acid–base interactions	Langmuir	Better removal of heavy metals in different pH concentration	[231]

Table 8. Cont.

Adsorbate	Adsorbent	Surface Area (m <sup>2</sup> /g)	Diameter (nm)	Q <sub>e</sub> /RE	Experimental Conditions			Removal Mechanism	Model	Comments	Ref.
					pH	IC	AL				
Zn(II)	Oxidized MWCNTs	-	-	3.83 mg/g	7	10–100 µg/L	25 mg/50 mL	Electrostatic interaction	Langmuir	Small diameter of CNTs removing greater amount of Hg(II) from aqueous solution	[246]
	Functionalized MWCNTs	250	10–25	2.42 mg/g	10	1.1 mg/L	0.09 g	Electrostatic interaction	Langmuir	Excellent potential for the removal of heavy metal ions	[247]
	Oxidized CNTs	-	-	74.63 mg/g	7	1200 mg/L	50 mg/25 mL	Chemical interaction	Langmuir	Economically feasible material with excellent heavy metal ion removal efficiency without any CNTs leakage	[235]
	Chitosan-MWCNTs	-	60–100	N/A	7	-	200 mg	Electrostatic interaction	N/A	The removal efficiency increases with increase of mass of both MWCNTs and Chitosan	[242]
	Nitrogen-doped magnetic carbon nanoparticles	-	-	9.31 mmol/g	8	12.82 mg/L	10 mg/500 mL	Chemical adsorption	Langmuir	Higher specific surface area and nitrogen make the nanomaterial an excellent adsorbent	[225]
Zn(II)	Oxidized MWCNTS	-	14	0.27 mmol/g	6.5–6.8	15 mg/L	5 mg/5 mL	Electrostatic interaction	Langmuir	Further research is necessary to understand the full mechanism	[199]

Table 8. Cont.

Adsorbate	Adsorbent	Surface Area (m <sup>2</sup> /g)	Diameter (nm)	Q <sub>e</sub> /RE	Experimental Conditions			Removal Mechanism	Model	Comments	Ref.
					pH	IC	AL				
	Sodium-hypochlorite-treated MWCNTs	-	<10	34.36 mg/g	-	60 mg/L	50 mg/100 mL	Electrostatic interaction	Langmuir	Zinc ion could be easily regenerated, and the adsorbent can be used for many cycles	[211]
	Sulfonated MWCNTs	28.7	-	43.16 mg/g	5	20 mg/L	25 mg/50 mL	Electrostatic interaction, surface complexation	D-R model	Enabling CNTs for wastewater treatment and composite formation or physical blending	[248]
	Magnetic MWCNTs	-	10–20	38.91 mg/g		30 mg/L	200 mg/1000 mL	Electrostatic interaction, physical interaction	-	Easily regenerate the Cu after removal from polluted water	[232]
Cu(II)	Oxidized CNT sheets	-	-	64.93 mg/g	7	200 mg/L	50 mg/25 mL	Chemical interaction	Langmuir	Considering the oxidize CNT sheets promising nanomaterial for heavy metal adsorption	[235]
	Chitosan/poly(vinyl) functionalized MWCNTs	-	5–20	11.1 mg/g	5.5	30 mg/L	0.5–2 wt%	Ion exchange	Langmuir	No loss in the adsorption capacity after four regeneration cycles	[249]
	MWCNTs modified with Chitosan	-	60–100	>95%	-	-	2000 mg	Electrostatic interaction	-	The removal efficiency increases with increase of mass of both MWCNTs and chitosan	[242]



Table 8. Cont.

Adsorbate	Adsorbent	Surface Area (m <sup>2</sup> /g)	Diameter (nm)	Q <sub>e</sub> /RE	Experimental Conditions			Removal Mechanism	Model	Comments	Ref.
					pH	IC	AL				
Co(II)	Chitosan-grafted MWCNTs	-	-	24.0 mg/g	-	10mg/L	1000 mg/1000 mL	N/A	-	Effective preconcentration and solidification of heavy metals in aqueous samples	[250]
	Poly(acrylic acid)-grafted MWCNTs	-	-	1.66 × 10 <sup>-4</sup> mol/g	6.8	1.69 × 10 <sup>-4</sup> mol/L	1.0 g/L	Surface complexation	Langmuir	Promising ability to use in water purification	[251]
	MWCNT <sub>S</sub> / iron oxide	-	-	0.18 mmol/g	6.4	4.2 mg/L	0.5 g/L	Ion exchange, surface complexation	Langmuir	Highlights the interaction between heavy metals and organic substances in wastewater	[252]
	Oxidized CNT sheets	-	-	85.74 mg/g	7	1200 mg/L	50 mg/25 mL	Chemical interaction	Langmuir	Considering the oxidized CNT sheets promising material for the removal of heavy metal ions	[235]
Ni(II)	HNO <sub>3</sub> -treated MWCNTs	102	10–20	17.86 mg/g	6.5	20 mg/L	0.8 g/L	Ion exchange	Langmuir	Better removal efficiency toward heavy metal ions	[253]
	MWCNTs modified with Chitosan	-	60–100	90%	-	-	2000 mg	Electrostatic interaction	-	The removal efficiency increases with an increase of mass of both MWCNTs and Chitosan	[242]
	Nitrogen-doped magnetic carbon nanoparticles	-	-	8.06 mmol/g	8	12.82 mg/L	10 mg/500 mL	Chemical adsorption	Langmuir	The removal efficiency was not very good for Ni compared to Cr	[225]

Table 8. Cont.

Adsorbate	Adsorbent	Surface Area (m <sup>2</sup> /g)	Diameter (nm)	Q <sub>e</sub> /RE	Experimental Conditions			Removal Mechanism	Model	Comments	Ref.
					pH	IC	AL				
	Poly(acrylic acid) (PAA)-oxidized MWCNTs	197	-	$6.615 \times 10^{-6}$ mol/g	5.4	5 mg/L	0.8 g/L	Electrostatic interaction, $\pi$ - $\pi$ interaction	Langmuir	Effective preconcentration and solidification of Ni(II) in liquid samples	[60]
	NaClO-modified SWCNTs	-	380	47.86 mg/g	7	10–80 mg/L	50 mg/100 mL	Electrostatic interaction	Langmuir	High removal affinity to heavy metals and can be used for water treatment	[216]
	MWCNTs/ Iron oxide	-	-	9.18 mg/g	-	6 mg/L	0.75 g/L	Ion exchange	Langmuir	Promising candidate for the solidification and preconcentration of heavy metal ions as well as for radionuclides from water	[201]
	Oxidized MWCNTs	-	5.5–14	49.26 mg/g	-	10–200 mg/L	20 mg/50 mL	Electrostatic interaction	-	Greater adsorption ability than raw MWCNTs in water	[254]
	Oxidized MWCNTs	197	10–30	>80%	8	6–20 mg/L	50 mg/200 mL	Electrostatic interaction	-	Excellent material for the adsorption of metal ions.	[255]
Ni(II)	MWCNTs	40–600	40–60	6.09 mg/g	7	25 mg/L	5 g/L	Ion exchange, surface complexation, chemical interaction	-	Excellent sorption of Ni <sup>+2</sup> ions with smaller equilibrium time	[256]

Table 8. Cont.

Adsorbate	Adsorbent	Surface Area (m <sup>2</sup> /g)	Diameter (nm)	Q <sub>e</sub> /RE	Experimental Conditions			Removal Mechanism	Model	Comments	Ref.
					pH	IC	AL				
U(II)	Diglycolamide-functionalized MWCNT (DGA-MWCNTs)	300–600	-	133.74 mg/g	7	-	1–10 mg	-	Langmuir	Adsorption efficiency increased with the increasing dose of adsorbent and temperature	[257]
Sr(II)	Oxidized-MWCNTs	-	-	36%	2–11	-	3 g/L	-	Diffuse layer model	Adsorption efficiency increased with increasing pH but decreased with the ionic strength	[258]
Eu (III)	Oxidized-MWCNTs	-	-	96%	2–11	-	3 g/L	-	Diffuse layer model	Higher adsorption efficiency for Eu(III) than Sr(II)	[258]

Q<sub>e</sub> = Max adsorption capacity (mg/g); RE = Removal efficiency; IC = Initial concentration; AL = Adsorbent loading; - = Data not available

## 6.2. Removal of Organics

Organics due to human actions [259], animals, or plant deterioration [260] are present in water and wastewater, usually in the form of dissolved and/or suspended particulate matters [135]. Zare et al. [261] reported that the adsorption efficiencies of CNTs for organic dyes can be enhanced after the functionalization. Methylene blue and orange were effectively removed from the water matrix by using oxidized MWCNTs [262,263] as compared to other types of modified adsorbents. Duman et al. [264] successfully used a novel nanocomposite, MWCNTs/carrageenan/ $\text{Fe}_3\text{O}_4$ , for effective removal of crystal violet. New nanocomposites can also be used for CNT modification to increase the adsorption capacity of cationic dyes. Sadegh and coworkers [265] reported that the adsorption of amide black can be significantly enhanced by using MWCNT-COOH-cysteamine. The results showed a maximum adsorption capacity of  $131 \text{ mg g}^{-1}$  for MWCNT-COOH-cysteamine, while it was  $90 \text{ mg g}^{-1}$  for MWCNT-COOH [265].

The adsorption capacities of CNT composites can be reduced effectively because of competition that occurred between several types of organic contaminants present in water and wastewater at the same time [266], similarly to as described for heavy metal adsorption. However, as reported by Ali et al. [135], the initial concentration did not have any significant effect on the removal capacities of CNTs for the removal of inorganic contaminants. As a result, customization and modifications of the surface of CNTs make them significant for selective adsorption of organic contaminants. Moreover, Wang et al. showed about 95% efficiency of the CNT composite for pharmaceutical and personal care product (PPCP) removal by increasing the specific surface area and aromatic ring of CNTs [267,268]. Jahangiri-Rad et al. [269] showed an adsorption capacity of  $496 \text{ mg g}^{-1}$  by using oxidized SWCNTs with large specific surface for the removal of blue 29 dye.

Adsorption capacities and mechanisms responsible for the interaction between organic contaminants and functionalized CNTs at different experimental conditions are described in Table 9. The increase in oxygen-containing functional groups on the surface of nanotubes results in a decrease in natural organic matters (NOM), because the  $\pi$ - $\pi$  interaction decreases due to more electrostatic repulsion [270]. The same type of results were shown in the case of higher pH [270]. Therefore,  $\pi$ - $\pi$  interactions are responsible between the large specific surface area of functionalized CNTs and NOM for the adsorption area of CNT [271,272]. Yang et al. [208] also reported similar results, as organics may be absorbed by  $\pi$ - $\pi$  interactions occurring between the surface of CNTs and aromatic rings of 1-naphthol.

CNTs have also been effectively used for the removal of pesticides [170]. Oxidized and as prepared MWCNTs showed excellent removal efficiency for diuron at  $\text{pH} \geq 7.0$  [273]. Oxidation of CNTs leads to the increase of their surface area, and pore size results in the higher adsorption of diuron [170]. Hamdi et al. [274] reported a reduction in uptake of chlordane and *p,p'*-dichlorodiphenyldichloroethylene from 78% to 23% in the roots of a lettuce crop by the addition of CNTs functionalized with amino group. The pesticides (1-pyrenebutyric, 2,4-dichlorophenoxyacetic, and diquat dibromide) were absorbed more (up to 70.6%) on semiconductor CNTs than metallic CNTs [275]. The lower density of electrons on the surface of semiconductor CNTs were found to facilitate higher adsorption [275]. On the other hand, during the batch modes, the pesticide removal was found to be limited [276]. SWCNTs and MWCNTs were used for the removal of diquat dibromide in the fixed bed system. The results were not higher than those of the batch system, but the pesticide absorbed completely in the fixed bed system by increasing the time [276].

The CNTs are also being used for nano-filtration of contaminants from aqueous solutions [162,277,278]. For adsorbents, the selectivity of nano-filters can be controlled by attaching different functional groups to the surface of CNTs [279]. Beside the hydrophobicity of CNTs, molecular dynamics simulations showed a weak interaction between water and nanotubes [280]. Hummer [281] explained that friction-free water flow was caused by the nano-scale pore size, which makes interaction energy smaller and finally lowers the interactions with water [281].

**Table 9.** Functionalized CNTs used for organic pollutants removal.

Adsorbent	Dye Pollutants	Surface area (m <sup>2</sup> /g)	Q (mg/g)	Removal percentage (%)	Optimum conditions	Remarks	Ref.
Oxidized SWCNT	Basic red 46 (BR 46)	400	49.45	-	pH 9, IC = 150 mg/L, AL = 0.05 g, Contact time = 100 min, 298 K,	Exothermic process favored at lower temperature range Orderly adsorption of dye due to negative entropy	[282]
HNO <sub>3</sub> -oxidized MWCNTs	Bromothymol blue (BTB)	96.8	55	97	pH 1, IC = 30 mg/L, AL = 0.02 g, T = 293.15 K,	Endothermic process of adsorption significantly affected by pH, initial concentration, sorbent dosage, and contact time	[262]
Functionalized CNT/Mg(Al)O	Congo red	148	1250	94	pH 7, AL = 30 mg, contact time = 75 min	Strong electrostatic interactions between dye particles and functional groups associated with the surface of nanomaterial	[283]
Magnetic MWCNTs-Fe <sub>3</sub> C nanocomposite	Direct red 23	38.7	85.5	-	pH 3.7, IC = 54 mg/L, AL = 0.04 g, T = 333 K,	Spontaneous endothermic adsorption process due to positive enthalpy	[284]
SWCNT-COOH	Malachite Green	400	22.33	-	pH 7, IC = 10 mg/L, 300 K,	Adsorption significantly affected by ionic strength, initial concentration, sorbent mass, contact time, and temperature More active functional groups on SWCNT-NH <sub>2</sub> adsorbed more dye than SWCNT-COOH	[285]
SWCNT-NH <sub>2</sub>	Malachite Green	400	29.36	-	pH 7, IC = 10 mg/L, T = 300 K,		
SWCNT-COOH	Methyl orange	400	25	-	pH 7, IC = 10 mg/L, T = 300 K,		
SWCNT-NH <sub>2</sub>	Methyl orange	400	27.15	-	pH 7, IC = 10 mg/L, T = 300 K,		
Oxidized MWCNTs	Methyl orange	165	10	-	AL = 20 mg/L, T = 313 K, stirring speed = 500 rpm	Initially, rapid adsorption was observed, but it slowed down with the time As the mixture temperature, agitation speed, and initial concentration increased, the adsorption efficiency also increased	[286]

Table 9. Cont.

Functionalized-CNTs loaded TiO <sub>2</sub>	Methyl orange	-	42.85	100	pH 6.5, IC = 5 mg/L, contact time = 30 min, T = 298 K	Highly active hydroxyl and amine functional groups made TiO <sub>2</sub> -CNT composite an effective adsorbent	[287]
Thiol-functionalized MWCNT (MWCNT-SH)	Methylene blue	400	166.67	-	pH 6, IC = 10 mg/L, AL = 20 mg, T = 298 K, Contact time = 60 min,	As the temperature and initial concentration increased, the adsorption efficiency also increased	[288]
Adsorbent	Phenol and its derivatives pollutants	Q (mg/g)	Surface area (m <sup>2</sup> /g)	Removal percentage (%)	Optimum conditions	Remarks	Ref.
KOH-modified MWCNTs	Bisphenol-A	0.20 mmol/g	494.48	-	pH 6, IC = 40 mg/L, contact time = 5 min, T = 298 K,	As the pH increased, the adsorption capacity decreased because of deprotonating; both the negatively charged function groups and adsorbates repel each other	[289]
HNO <sub>3</sub> -modified MWCNTs	Bisphenol-A	0.59 mmol/g	153.79	-	pH 6, IC = 40 mg/L, contact time = 30 min, T = 298 K,		
SOCl <sub>2</sub> /NH <sub>4</sub> OH-modified CNT	Bisphenol A	69.93	94.8	-	pH 6.5, IC = 10 mg/L, AL = 0.125 g/L, T = 280 K,	Adsorption efficiency increased with the initial concentrations	[290]
NH <sub>3</sub> -treated MWCNTs	Chlorophenols (CP)	110.3	195	-	pH 3.8, T = 298 K,	The adsorption capacity increased due to higher pores size, $\pi$ - $\pi$ interactions, and hydrophobicity of nanocomposite Effective nanomaterial with smaller equilibrium time	[291]
HNO <sub>3</sub> and KMnO <sub>4</sub> -Functionalized MWCNTs	Phenol	76.92	-	88	IC = 500 mg/L, Agitation speed = 200 rpm, T = 298 K	Adsorption capacity can be greatly affected by pH and adsorbent mass	[292]
Oxidized SWNTs	<i>p</i> -Nitrophenol (PNP)	206	-	97.9	IC = 0.01 mg/mL, agitation time = 30 min T = 293 K	Open ends of nanotubes, higher surface area, and the functional groups (hydroxyl, carbonyl, and carboxyl) were responsible for higher adsorption	[293]

Table 9. Cont.

Nitrogen-doped carbon nanotubes (CNx)	Phenol	0.16 mmol/g	102	-	pH 7, AL = 0.6mmol/L, 298 K,	$\pi$ - $\pi$ interaction occurred between the functional groups and phenol; more oxidized CNTs adsorbed less phenol	[294]
MWCNT-COOH	Phenol	0.15 mmol/g	-	-	IC = 0.417 mg/L, AL = 10 mg, T = 293 K	Higher adsorption of CP than phenol resulted because of the different solubility of these contaminants	[295]
	3-Chlorophenol (CP)	0.37 mmol/g	-	95	IC = 1.25 mg/L, AL = 10 mg, T = 293 K		
Acid-functionalized MWCNT (MWCNT-COOH)	2-Nitrophenol	256.41	197.83	-	pH 5.5, IC = 45 mg/L, T = 298 K,	Excellent adsorbent due to strong interactions between 2-Nitrophenol and surface functional groups	[296]

Q = adsorption capacity, IC = Initial concentration, AL = Adsorbent loading and T = Temperature.



### 6.3. Removal of Microorganisms

Bacteriological contaminants, deteriorating the assimilative capacity of water bodies leading to diverse impacts on surrounding environment, are a major challenge [297]. Bacteriological contaminants are often found in surface waters, wastewaters, and respective treatment plants [145,298]. CNTs, with their diverse range of surface and functional characteristics, have a high-affinity interaction with biological contaminants. CNTs have been proven for their higher adsorption capacities, inactivation efficiencies of viral or bacterial spores, and greater antimicrobial potential than conventional sorbents because of their greater surface area [170,299,300].

Previous studies that have reported that CNTs can inactivate or remove a variety of microorganisms, including bacteria, are shown in Table 10. The inactivation of *E. coli* was observed by using SWCNTs because of its penetration into the bacterial cell wall [301]. In addition, surface-modified CNTs with different chemical groups destroy the cell wall of microorganisms more strongly than the raw and polymer-grafting CNT membranes [135]. Furthermore, the straight interaction of microbes (i.e., *E. coli*) and functionalized CNTs can cause adverse effects on the metabolisms and morphological structure of the cell wall of bacteria [135]. The penetration of CNTs into the cell wall of microorganisms is the main reason for their higher inactivation affinity [301].

CNTs functionalized with silver nanoparticles showed excellent ability to inactivate microorganisms. For example, Ihsanullah and his coworkers [302] synthesized and successfully used silver-doped CNTs for the inactivation of *E. coli* bacteria. The results proved that 100% of the bacteria were killed due to toxicity of synthesized silver doped nanomaterial [302]. In addition, many researchers reported that the diameter of nanotubes can be an important factor of concern for inactivation pathogenic microorganisms (Table 10) in water and wastewater.

The single kinetics of CNTs also reveal the CNTs' ability to eliminate pathogens in water and wastewater treatment, and microbes remain on the CNT surface based on deep filtration mechanisms [303,304]. Brady and coworkers [299] used a poly vinylidene fluoride-based SWCNT filter for excellent removal of *E. coli* bacteria at low pressure. The results show that the cells were completely captured by the filter [299]. In addition to filtrations, an excellent removal efficiency of MS2 virus was observed by filtering the sample through a controlled nano porous CNT based filter at a pressure of about 8 to 11 bar [304]. Beside the filtrations, CNTs are widely used in water and wastewater treatment as an antimicrobial agent, as described earlier. This behavior makes them a replacement for chemical disinfectants as a new method for controlling pathogens [301,305–308]. The applications of CNTs for disinfection treatment in water and wastewater avoid the materialization of unsafe disinfectant by-products like trihalomethanes, aldehydes, and haloacetic acids due to their low oxidation state and solubility in water. Therefore, it is necessary to promote the dispersibility of these compounds; a surfactant or a polymer such as sodium dodecylbenzenesulfonate, polyvinylpyrrolidone is used [309]. Due to the excellent mechanical properties on CNTs, they act as scaffolds for antimicrobial agents, such as silver nanoparticles [310,311] and antibacterial lysozyme [307,309].

Table 10. Functionalized CNTs used for disinfection.

Contaminants	Adsorbents Types	AL	IC	RE	Removal Mechanism	Comments	Ref
<i>Escherichia coli</i> ( <i>E. coli</i> )	Silver-doped CNT membrane	-	$1 \times 10^6$ CFU/mL	100%	-	All the bacteria were inactivated by membrane with 10% silver loadings in 60 min only	[312]
	Silver-nanoparticle-loaded CNTs	2.5 µg/mL	$10^6$ CFU/mL	89%	-	Effectively inactivate the pathogen from wastewater effluents, resistance toward bacterial adhesion	[313]
	Chitosan/CNT nanocomposites	2 wt%	$1.5 \times 10^8$ to $5.0 \times 10^8$ CFU/mL	2.89 log reduction	Physical interaction and surface complexation	Higher antimicrobial activity at the low contact time (10 min) and low concentration (1%)	[314]
	Acidic-conditioned MWCNTs	200 µg/mL	$10^6$ to $10^9$ CFU/mL	-	Steric obstruction	Inactivation of pathogens was due to both MWCNT functionalization and nutrition level	[315]
	1-octadecanol-functionalized MWCNTs	0.2 g/100 mL	$3.5 \times 10^7$ CFU/mL	100%	Polarization	The interaction of microwaves with f-CNTs is an innovative approach that has the potential to be employed for water disinfection	[316]
<i>Staphylococcus aureus</i>	CNT–Ag nanohybrid	2.5 µg/mL	$10^6$ CFU/mL	100%	-	Effectively inactivate the pathogen from wastewater effluents, resistance toward bacterial adhesion	[313]
	Chitosan/CNTs nanocomposites	2 wt%	$1.5 \times 10^8$ to $5.0 \times 10^8$ CFU/mL	4.9 log reduction	Physical interaction and surface complexation	Higher antimicrobial activity at the low contact time (10 min) and low concentration (1%)	[314]
<i>Aspergillus flavus</i>	Chitosan/CNTs nanocomposites	2 wt%	$1.5 \times 10^8$ to $5.0 \times 10^8$ CFU/mL	5.5 log reduction	Physical interaction and surface complexation	Higher antimicrobial activity at the low contact time (10 min) and low concentration (1%)	[314]

RE = removal efficiency (%) / log reduction; AL = adsorbent loading; IC = initial concentration, mg/L; CFU/mL = colony-forming unit per milliliter.

## 7. Conclusions

Purifying water from assorted contaminants is challenging, and carbon nanotube-based nanocomposites can provide simple as well as effective water decontamination/disinfection. This review shows that functionalized CNTs are a new generation of pollution management materials. These materials have excellent adsorption capacities and work effectively in removing organics, inorganics, and biological species. Various sorption mechanisms include physical adsorption, electrostatic interaction, surface complexation, and chemical interactions between surface functional groups and metal ions. The effects of pH, CNT dosage, time, ionic strength, temperature, and surface charge on the adsorption of heavy metal ions on carbon nanotube surface were also discussed. Even functionalized CNTs also has antibacterial efficacy against Gram-positive and Gram-negative bacteria.

Almost all the studies show effective removal of contaminants and have been performed using deionized water, but the potential of functionalized CNT nanomaterials needs to be verified under real water conditions. In wastewater, carbonates, phosphates, and silicates can successfully compete with target metals/organics for adsorption sites in nanostructures. Other important components of source water and wastewater are natural organic materials, such as fulvic acid and humic acid, which will occupy the surface of CNTs and thus affect the adsorption of contaminants on the nanostructures. The effectiveness of CNT-based nanotechnology should be evaluated under real water conditions.

Many researchers have also focused on evaluating the adsorption–desorption capacity to make it a cost-effective adsorbent for use in wastewater treatment. However, more studies are encouraged to check the feasibility of reuse.

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